



## ORIGINAL ARTICLE

# Exploring Health Risks Associated with Heavy Metal Contamination in Groundwater from Industrial Zones in Samut Prakan Province, Thailand

Somkid Tangan<sup>1</sup>, Cherlyn Sirisetpop<sup>\*2</sup>, Potchanun Sriphothong<sup>3</sup>

<sup>1</sup>Faculty of Science and Technology, Valaya Alongkorn Rajabhat University under the Royal Patronage, Pathumthani, 13180, Thailand

<sup>2</sup>Faculty of Public Health, Valaya Alongkorn Rajabhat University under the Royal Patronage, Pathumthani, 13180, Thailand

<sup>3</sup>School of Engineering Science and Technology, Sarasas Suvarnabhumi Institute of Technology, Samut Prakan, 10540, Thailand

(Received: 30 January 2025

Accepted: 12 March 2025)

## KEYWORDS

Health risk assessment;  
Contamination;  
Heavy metals;  
Groundwater

**ABSTRACT:** Groundwater in industrial zones is often contaminated with heavy metals, posing significant health risks. This study investigates the contamination of arsenic, nickel, lead, and zinc in groundwater from the Bangkok, Phra Pradaeng, Nakhon Luang, and Nonthaburi aquifers and assesses the health risks associated with heavy metal contamination in groundwater within Samut Prakan Province. Groundwater samples were collected from observation wells (n=16) and analyzed for dissolved heavy metals using USEPA 200.7 and 6010D standard methods. Health risks were assessed for both non-carcinogenic and carcinogenic effects using USEPA models. The results showed that arsenic levels exceeded USEPA and WHO standards in the Nakhon Luang and Nonthaburi aquifers. Nickel and lead levels surpassed the permissible limits of USEPA, WHO, and NOAA in all aquifers. Exposure assessments indicated that children had higher levels of heavy metals, with ingestion as the main route of exposure. Non-carcinogenic health risks from arsenic, nickel, lead, and zinc through ingestion and dermal contact were found to be significant for both children and adults ( $HI > 1$ ) in all groundwater aquifers. The carcinogenic health risk assessment revealed unacceptable cancer risks from the ingestion of arsenic, nickel, and lead, and dermal exposure to arsenic and nickel ( $TCR > 10^{-4}$ ) in all groundwater aquifers. These findings indicate the need for close monitoring and management of industrial activities to prevent further heavy metal contamination in groundwater, thereby reducing health risks for individuals relying on this resource.

## INTRODUCTION

Heavy metals are chemical substances that can significantly impact environmental quality [1] and are often toxic to living organisms. Poor environmental quality due to heavy metal contamination can result in harmful effects on humans, animals, and plants [2]. For example, mercury exposure can cause Minamata disease

in humans, arsenic exposure can lead to neurological disorders, nickel contamination can impair blood circulation and slow plant growth [3], while lead exposure can cause disorders in both the human excretory system and the process of photosynthesis in plants [4]. In addition to toxic heavy metals, there is

\*Corresponding author: cherlyn.si@vru.ac.th (Ch. Sirisetpop)  
DOI: 10.60829/jchr.2025.1198062

another group of heavy metals that are essential for living organisms. These metals are necessary for various physiological processes in the body to prevent disease. However, excessive intake of these metals can cause health issues and diseases if consumed in excessive quantities [5]. For instance, excessive zinc intake can negatively affect the nervous and immune systems [6]. This demonstrates that both toxic metals and those necessary for life can have harmful effects on living organisms when accumulated in amounts the body cannot eliminate, leading to the onset of diseases.

Environmental media such as soil, water, and air [7] are commonly impacted by heavy metal contamination. Sources of heavy metals can be both natural and anthropogenic [8], including industrial activities that use heavy metals, such as those in the paint, metal, and battery industries. According to the Department of Industrial Works in Thailand [9], there were 73,232 factories in the country in 2022, many of which are associated with the use of heavy metals. These include chemical manufacturing, pesticide production, plastic manufacturing, metal sheet, electrical appliances, and automobile and vehicle part manufacturing. Typically, these factories are located within various industrial estates across Thailand, with 6,958 factories situated in Samut Prakan province. As a result, Samut Prakan is one of the provinces at high risk of heavy metal contamination from industrial activities. One major environmental concern in this area is groundwater contamination with heavy metals. This contamination can spread to various regions and poses significant challenges in treatment before the water is used by humans.

In Samut Prakan industrial area, the primary causes of heavy metal contamination in groundwater are industrial activities that lead to the leakage of heavy metals and the leaching of contaminated soil into the groundwater [10-12]. Heavy metal leaching into groundwater may result from rainfall or activities involving significant water leakage into the environment, such as the use of water [11-13] and foam for firefighting in the severe fire at the Mingti chemical plant. This event involved 18 hours of water and foam usage, which may have contributed to heavy metal contamination in the groundwater through the leaching of metals from the soil into the aquifer [14].

Factories in Samut Prakan industrial zones that use heavy metals include battery plants, automobile parts factories, chemical manufacturing plants, metal parts factories, electroplating facilities, and pesticide manufacturing plants [15]. The heavy metals commonly used in industrial processes include arsenic, nickel, lead, and zinc. Therefore, industrial areas in these zones are at risk of heavy metal leakage, which could affect the quality of groundwater.

Studies on heavy metal contamination in groundwater and its effects on human health have shown significant concerns in industrial zones, such as Peenya in Bengaluru, India, and Lahore in Pakistan. In these areas, heavy metal contamination has led to substantial health risks for individuals who rely on groundwater for daily use [16]. In Pakistan, the contamination has resulted in severe health issues, with an approximate 30% death rate among those affected by water-borne diseases and heavy metals [17].

The spread of heavy metals in groundwater varies by region and is influenced by several factors, including hydrogeological processes, bedrock composition [18], the adsorption of heavy metals in aquifers [19], sedimentation, and the amount of metal leakage from sources. Additionally, the mobility of heavy metals in groundwater may be affected by chemical reactions with dissolved substances, such as high salt concentrations. If groundwater contains high levels of salts, heavy metals can form soluble compounds that spread more easily [20]. In the Samut Prakan industrial area, the presence of both heavy metal sources and high salt concentrations in groundwater, due to seawater intrusion, further exacerbates the spread of contamination. Thus, if industrial activities in this area lead to heavy metal leakage, the spread of contamination in the groundwater will intensify, potentially affecting human health even in areas distant from the original contamination sources. Therefore, due to the aforementioned reasons, the analysis of heavy metal contamination in groundwater remains a critical issue for further investigation. Additionally, this study primarily aims to assess the health risks associated with the exploitation of heavy metal-contaminated groundwater, a topic that has not been previously explored in this study area. The findings will provide valuable guidance for government agencies

responsible for pollution control, as well as for the public engaged in groundwater exploitation, ensuring its proper management to mitigate potential health risks in the future.

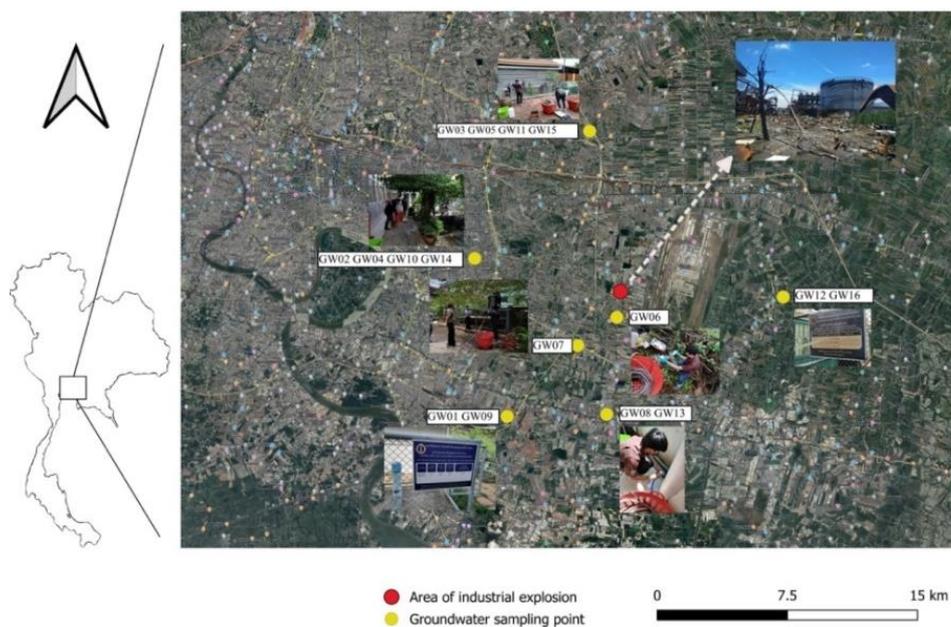
The aim of this study is to assess the levels of heavy metal contamination—specifically arsenic, nickel, lead, and zinc—in groundwater from the Samut Prakan industrial area. The study focuses on contamination of heavy metals in the Bangkok, Phra Pradaeng, Nakhon Luang, and Nonthaburi aquifers. Additionally, this study aims to evaluate the associated health risks from exposure to these metals within the Samut Prakan industrial area.

## MATERIALS AND METHODS

### *Study area and groundwater sampling*

The groundwater sampling area is located within a 30-kilometer radius around the Samut Prakan industrial area,

as shown in Figure 1. The coordinates of the sampling locations are provided in Table 1. Samples were collected from observation wells across four aquifers: Bangkok (3 wells, n=3), Phra Pradaeng (3 wells, n=3), Nakhon Luang (6 wells, n=6), and Nonthaburi (4 wells, n=4), at average depths of 50, 100, 150, and 200 m. [21], respectively. Groundwater sampling was conducted in collaboration with the Department of Groundwater Resources, Thailand, in August 2021, about one month after the severe fire at Mingti Chemical. Groundwater was pumped from the observation wells using a pump system, in accordance with the Department's standard procedure. At each observation well, three samples were collected. The volume of each sample was 2 L. To preserve the samples, concentrated nitric acid (QReC, Newzealand) was added to lower the pH to below 2 during transport. The samples were stored at 4°C [22] in a light-free environment until reaching the laboratory.



**Figure 1.** Map of groundwater sampling locations generated using QGIS.

**Table 1.** Geographic coordinates of the groundwater sampling points, groundwater depth levels, and sampling locations

Aquifer	Sample No.	GPS coordination		Depth (m)	Location
		East	North		
Bangkok	GW01	100°39'06.5"E	13°36'28.8"N	54	Theparak local park
	GW02	100°38'06.1"E	13°41'27.3"N	50	Wat wachiratham sathit
	GW03	100°41'48.4"E	13°45'26.0"N	53	Wat lad bua khao
Phra Pradaeng	GW04	100°38'06.1"E	13°41'27.3"N	87	Wat wachiratham sathit
	GW05	100°41'48.4"E	13°45'26.0"N	99	Wat lad bua khao
	GW06	100°42'38.1"E	13°39'34.5"N	116	Wat khlong chuat lak khao
	GW07	100°41'22.9"E	13°38'41.7"N	140	Bang phli telephone exchange
Nakhon luang	GW08	100°42'17.3"E	13°36'31.4"N	163	Bang phli yai klang temple
	GW09	100°39'06.5"E	13°36'28.8"N	130	Theparak local park
	GW10	100°38'06.1"E	13°41'27.3"N	181	Wat wachiratham sathit
	GW11	100°41'48.4"E	13°45'26.0"N	160	Wat lad bua khao
	GW12	100°47'57.5"E	13°40'10.5"N	146	Wat sriwaree noi
Nonthaburi	GW13	100°42'17.3"E	13°36'31.4"N	211	Bang phli yai klang temple
	GW14	100°38'06.1"E	13°41'27.3"N	217	Wat wachiratham sathit
	GW15	100°41'48.4"E	13°45'26.0"N	204	Wat lad bua khao
	GW16	100°47'57.5"E	13°40'10.5"N	203	Wat sriwaree noi

### Sample preparation and heavy metal analysis

Sample preparation for the analysis of dissolved heavy metals in contaminated groundwater was carried out in accordance with USEPA standard Method 200.7 [23,24]. A 50 ml. of groundwater sample was filtered using a filter paper (Whatman, USA) with a porosity of 0.45  $\mu\text{m}$ . The filtrate was then analyzed for heavy metals using Inductively Coupled Plasma Atomic Emission Spectroscopy (ICP-AES) (PerkinElmer model 8300, USA) following USEPA Standard Method 6010D [25]. The wavelengths used for the analysis of arsenic (As), nickel (Ni), lead (Pb), and zinc (Zn) were 193.693, 231.604, 220.353, and 213.856 nm., respectively.

### Health risk assessment

#### Exposure assessment

The health risk assessment from heavy metal exposure through ingestion and dermal contact was conducted following the USEPA guidelines [26], as outlined in Equations (1-2).

$$\text{EDI}_{\text{ing}} (\text{mg kg}^{-1} \text{ day}^{-1}) = (C \times \text{IngR} \times \text{EF} \times \text{ED}) / (\text{BW} \times \text{AT}) \quad (1)$$

$$\text{EDI}_{\text{der}} (\text{mg kg}^{-1} \text{ day}^{-1}) = (C \times \text{SA} \times \text{Kp} \times \text{ET} \times \text{EF} \times \text{ED} \times \text{CF}) / (\text{BW} \times \text{AT}) \quad (2)$$

Where, EDI represents the amount of heavy metals entering the body via ingestion ( $\text{EDI}_{\text{ing}}$ ) and dermal contact ( $\text{EDI}_{\text{der}}$ ); C is the concentration of heavy metals in groundwater ( $\text{mg L}^{-1}$ ); IngR is the ingestion rate of groundwater (1.8 L per day for children, 2.2 L per day for adults) [27]; EF is the exposure frequency (365 days per year) [27]; ED is the exposure duration (6 years for children, 70 years for adults) [28]; BW is body weight (15 kg for children, 70 kg for adults) [28]; AT is the average exposure time ( $\text{AT} = \text{ED} \times 365$  days) [29]; Kp is the dermal permeability coefficient in water ( $0.001 \text{ cm hr}^{-1}$  for As,  $2 \times 10^{-4} \text{ cm hr}^{-1}$  for Ni,  $4 \times 10^{-3} \text{ cm hr}^{-1}$  for Pb,  $6 \times 10^{-4} \text{ cm hr}^{-1}$  for Zn) [27, 30]; SA is the skin surface area ( $6,600 \text{ cm}^2$  for children,  $18,000 \text{ cm}^2$  for adults) [27]; ET is the exposure time per day (1 hour per day for children, 0.58 hour per day for adults) [27]; and CF is the conversion factor ( $10^{-3} \text{ L cm}^{-3}$ ) [31].

### ***Non-Carcinogenic health risk assessment***

The non-carcinogenic health risk assessment was conducted by calculating the Hazard Quotient (HQ), which is a numerical evaluation of the ratio between the amount of each heavy metal entering the body and its reference dose, as expressed in the Equation (3). Additionally, the Hazard Index (HI) was calculated to evaluate the risk from multi-element exposure via multiple pathways, as shown in Equations (4).

$$HQ = EDI/RefD \quad (3)$$

$$HI = \sum_{i=1}^n HQ_i \quad (4)$$

Where, HQ represents the hazard quotient (unitless); HI represents the hazard index (unitless); RefD denotes the reference dose; and *i* refers to the metal species evaluated, including As, Ni, Pb, and Zn. The reference dose values for As, Ni, Pb, and Zn from dermal exposure are  $1.23 \times 10^{-4}$ ,  $5.4 \times 10^{-3}$ ,  $5.24 \times 10^{-4}$ , and  $6 \times 10^{-2}$  [32], respectively. The reference dose values for As, Ni, Pb, and Zn from ingestion are  $3 \times 10^{-4}$ ,  $2 \times 10^{-2}$ ,  $1.4 \times 10^{-3}$ , and  $3 \times 10^{-1}$  [32], respectively.

For the non-carcinogenic health risk assessment based on the HQ and HI values, as outlined by the USEPA, if the calculated HQ or HI values exceed 1, it indicates a potential health risk from heavy metal exposure. Conversely, if the HQ or HI values are less than 1, it implies there is no significant risk of non-carcinogenic health effects from the exposure to heavy metals.

### ***Carcinogenic health risk assessment***

The carcinogenic health risk assessment is a numerical evaluation method used to estimate the probability of cancer development following exposure to carcinogenic substances through various pathways. In this study, the health risks associated with exposure to As, Ni, and Pb through ingestion and As and Ni through dermal contact were assessed. The individual health risk from each heavy metal was calculated using the Cancer Risk (CR) as shown in Equation (5), while the health risk from multi-element exposure via multiple exposure pathways was determined by calculating the Total Cancer Risk (TCR) as shown in Equation (6).

$$CR = EDI \times SF \quad (5)$$

$$TCR = \sum_{i=1}^n CR_i \quad (6)$$

Where, CR is carcinogenic risk (unitless); TCR is total carcinogenic risk (unitless); SF is the slope factor; and *i* refers to the metal species evaluated, including As, Ni and Pb. The slope factors for As, Ni, and Pb for ingestion are 1.5, 1.7, and 0.0085 [32], respectively, while the slope factors for As and Ni for dermal exposure are 3.66 and 42.5 [32], respectively.

The health risk assessment for cancer development from exposure to heavy metals, based on the CR and TCR values provided by the USEPA [26], indicates that if the CR or TCR value is less than  $10^{-6}$ , there is no risk of cancer from heavy metal exposure. However, if the CR or TCR value exceeds  $10^{-6}$ , there is a risk of cancer development. If the CR or TCR value falls between  $10^{-6}$  and  $10^{-4}$ , the risk is considered acceptable, but if the CR or TCR value exceeds  $10^{-4}$ , the risk is deemed unacceptable [33, 34].

### ***Statistical analysis***

The statistical analysis was conducted using SPSS software. Descriptive statistics were presented as means and standard deviations (SD), while inferential statistics were derived from two-way ANOVA and Pearson's correlation coefficient. A two-way ANOVA was conducted to analyze the variations in heavy metal concentrations across different groundwater aquifer levels. The analysis considered three factors: metal species (F1), groundwater aquifer levels (F2), and the interaction between these two factors ( $F1 \times F2$ ), with a significance level set at 0.05. Pearson's correlation coefficient was used to analyze the sources of heavy metal contamination in groundwater, with a significance level of 0.05 [35-37].

### ***Quality control and quality assurance***

Quality control throughout the analysis process was carried out from sample collection, sample preservation, sample extraction, to sample analysis according to standard procedures. The chemicals used were AR grade. Glassware used for sample extraction was decontaminated by soaking in 10% nitric acid for 48 hours, followed by cleaning with DI water and drying at

120°C for 4 hours [38] before analysis. A standard curve was prepared at concentrations of 0, 0.5, 1, 2, and 5 mg L<sup>-1</sup>. The results showed that the correlation coefficient ( $r^2$ ) was greater than 0.995, which is within the acceptable range ( $r^2 > 0.995$ ) [39]. Additionally, quality control was performed by determining the % recovery using standard reference material (PerkinElmer, Lot: 3-18mkby1, USA), which ranged from 97.17-108.74%, within the acceptable range of 80-120% [40,41]. Sample analysis was performed with triplicate measurements [42], and the results had to show a %RSD < 20% [43].

## RESULTS AND DISCUSSION

### *Contamination of heavy metals in groundwater*

The contamination of heavy metals in groundwater (in mg L<sup>-1</sup>) is presented in Table 2. The ranges of contamination for As, Ni, Pb, and Zn in groundwater are ND-0.041, 0.646-1.931, 0.026-0.107, and 0.027-6.450 mg L<sup>-1</sup>, respectively. The average concentrations of As (0.015), Ni (1.496), Pb (0.087), and Zn (1.431) were highest in the groundwater aquifers from Nonthaburi, Phra Pradaeng, Bangkok, and Nakhon Luang, respectively. Meanwhile, the groundwater aquifers in Bangkok showed the lowest average concentrations of As (ND) and Zn (0.063), while Ni (0.976) and Pb (0.055) were lowest in the Nonthaburi aquifers. The factor that significantly influenced the levels of heavy metal contamination in the groundwater aquifers was the type of heavy metal (Factor 1) at a significance level of 0.01 (Table 1). On the other hand, the factor of groundwater aquifer (Factor 2) and the interaction between Factor 1 and Factor 2 had no effect on the heavy metal concentrations in the groundwater aquifers.

The differences in concentrations of heavy metals in the various groundwater aquifers can be attributed to the specific characteristics of the metals themselves, such as principal hydrochemical processes, interactions of heavy metals with soil and rock, and various geological formations. In addition, external factors affecting the dissolution of heavy metals in groundwater, such as pH value, temperature, groundwater flow rate, oxidation reactions [10], geological conditions [44], and the effect of parent rock and bedrock [44], can lead to varying dissolution capacities of the heavy metals.

A comparison of heavy metal contamination in groundwater with the maximum contamination levels in groundwater established by USEPA and NOAA and the drinking water standards defined by WHO, demonstrated that the concentrations of Ni and Pb in the groundwater samples (GW01–GW16) and the average concentrations of Ni and Pb in each groundwater aquifer exceeded the standards of WHO, USEPA, and NOAA. In contrast, the average concentrations of As and Zn from all groundwater aquifers were below the standards. However, the contamination levels of As and Zn in some groundwater samples exceeded the standards. Specifically, As concentrations exceeding the USEPA and WHO standards in the Nakhon Luang aquifer (GW07-GW09 and GW12) and Nonthaburi groundwater aquifer (GW14 and GW16). Additionally, Zn concentrations exceeded the NOAA, USEPA, and WHO standards in the Nakhon Luang aquifer (GW07). These findings suggest that the contamination of heavy metals in groundwater in the study area is not suitable for consumption due to the potential health risks posed by high levels of Ni and Pb.

**Table 2.** Concentration of heavy metals (mg L<sup>-1</sup>) in groundwater in the presented study, with standard limits, statistical analysis results using two-way ANOVA, and comparisons with other previous studies

Aquifer	Sample No.	Concentration of heavy metals (mg L <sup>-1</sup> ) in groundwater (mean±SD)				Reference
		in presented study				
		As	Ni	Pb	Zn	
Bangkok	GW01	ND	1.480±0.48	0.107±0.03	0.093±0.02	
	GW02	ND	1.783±0.02	0.046±0.02	0.041±0.01	
	GW03	ND	1.014±0.32	0.107±0.02	0.055±0.01	
	Average	ND	1.426	0.087	0.063	
Phra Pradaeng	GW04	ND	1.698±0.01	0.078±0.01	0.027±0.01	
	GW05	0.003±0.03	1.724±0.01	0.066±0.02	2.969±0.10	
	GW06	ND	1.065±0.44	0.080±0.03	0.378±0.03	
	Average	0.001	1.496	0.075	1.125	
Nakhon Luang	GW07	0.012±0.12	0.909±0.00	0.072±0.03	6.450±0.01	
	GW08	0.013±0.00	1.371±0.48	0.101±0.03	0.596±0.08	
	GW09	0.023±0.02	1.522±0.52	0.079±0.03	0.118±0.01	
	GW10	ND	1.931±0.02	0.063±0.02	0.733±0.04	
	GW11	0.002±0.03	0.726±0.01	0.058±0.03	0.628±0.02	
	GW12	0.034±0.02	1.216±0.01	0.081±0.02	0.061±0.01	
	Average	0.014	1.279	0.076	1.431	
	GW13	ND	0.689±0.01	0.026±0.02	0.323±0.01	
Nonthaburi	GW14	0.017±0.03	1.556±0.01	0.067±0.03	0.188±0.02	
	GW15	ND	0.646±0.28	0.071±0.02	0.394±0.01	
	GW16	0.041±0.01	1.014±0.32	0.055±0.02	0.138±0.01	
	Average	0.015	0.976	0.055	0.261	
Average concentration in present study		0.009	1.272	0.072	0.825	
<b>Standard limits for heavy metal concentrations (mg L<sup>-1</sup>) in groundwater.</b>						
USEPA standard		0.01	0.1	0.015	5	[46,47]
WHO standard		0.01	0.07	0.01	3	[48]
NOAA standard		0.05	0.1	0.015	5	[49]
<b>Concentrations of heavy metals (mg L<sup>-1</sup>) in groundwater from studies conducted in other areas.</b>						
Peenya Industrial Area, India		0.012	0.014	0.009	0.126	[16]
Pepper production area, China			0.022		0.065	[50]
Hayatabad Industrial Estate, Pakistan			4.49	5.86	44.9	[51]
Dar es Salaam city, Tanzania		6.64		0.02	1.19	[52]
Industrial zones, Southern India			0.4	0.8	0.54	[53]
Industrial area in Sheikhpura district, Pakistan		38.49 × 10 <sup>-3</sup>	0.03 × 10 <sup>-3</sup>	0.05 × 10 <sup>-3</sup>	1.07 × 10 <sup>-3</sup>	[54]
Meghna Ghat industrial area, Bangladesh				0.1819		[55]
<b>Two-way ANOVA test</b>						
Source of variation		F-value		P value		
Factor 1 (type of heavy metal)		7.590		0.000*		
Factor 2 (aquifer)		0.766		0.517		
Interaction		0.609		0.827		

ND, not detected; \*, significant at 0.01 level

When comparing the heavy metal contamination in the study area's groundwater with other areas with similar

sources from industrial factories and human activities (Table 2), it was found that the average concentration of

As in groundwater in the Samut Prakan industrial area was higher than in the industrial area in Sheikhpura District (Pakistan). The concentrations of Ni and Zn in groundwater in Samut Prakan were higher than in the Peenya Industrial Area (India), the industrial zones in southern India, pepper production area in China and the industrial area in Sheikhpura District (Pakistan). The concentration of Pb in groundwater in Samut Prakan was also higher than in the Peenya Industrial Area (India),

Dar es Salaam city (Tanzania), and the industrial area in Sheikhpura District (Pakistan). The higher concentrations of certain heavy metals in the study area compared to other areas may be due to the greater number of sources of heavy metals in the study area and the better prevention systems for heavy metal leakage in the other areas. To reduce the heavy metal contamination in the study area, improvements in the prevention systems for leakage should be made.

**Table 3.** Pearson correlation coefficient values between heavy metal concentrations in groundwater from different aquifers

Aquifer	Element	Arsenic	Nickel	Lead	Zinc
Bangkok	Arsenic	-	-	-	-
	Nickel		1	-0.799	-0.141
	Lead			1	0.708
	Zinc				1
Phra Pradeang	Arsenic	1	0.530	-0.991	0.994
	Nickel		1	-0.637	0.434
	Lead			1	-0.971
	Zinc				1
Nakhon Luang	Arsenic	1	-0.033	0.517	-0.182
	Nickel		1	0.171	-0.406
	Lead			1	-0.149
	Zinc				1
Nonthaburi	Arsenic	1	0.467	0.185	-0.911
	Nickel		1	0.390	-0.729
	Lead			1	-0.076
	Zinc				1

The analysis of the sources of heavy metal contamination in groundwater, based on correlation calculations shown in Table 3, found that the metals had no significant correlation at the 0.05 significance level. This indicates that the sources of heavy metals contributing to the contamination of groundwater in the study area are not correlated [56,57], meaning it is not possible to identify a single source for the heavy metals. Heavy metal contamination in the groundwater of the study area may be attributed to multiple sources or originate from distinct origins. This is consistent with the nature of the study area, as there are various types of industrial factories that can contribute to the contamination of heavy metals, such as chemical factories, metal wire manufacturing, electrical circuit board production, ZnCl<sub>2</sub>, ZnSO<sub>4</sub>, and ZnCO<sub>3</sub> chemical plants, can manufacturing factories, paint factories, pesticide and insecticide

factories, automotive parts factories, fertilizer factories, battery factories, hazardous waste disposal sites in industrial parks, industrial waste disposal and treatment factories, galvanizing and metal coating factories, and leather tanning factories. Furthermore, more than 350 factories are located in the surrounding industrial areas, some of which, both within the industrial parks and nearby areas, may use multiple types of heavy metals (As, Ni, Pb, and Zn) in their industrial processes, which could contribute to contamination if leakage occurs into the groundwater.

#### *Exposure assessment*

Health risk assessment, encompassing both non-carcinogenic and carcinogenic risks from heavy metal exposure through ingestion and dermal contact, involves

calculating the quantity of substances entering the body by determining the EDI values for As, Ni, Pb, and Zn, as shown in Table 4.

The results of the metal intake calculations indicate that heavy metals primarily enter the body via ingestion (EDI<sub>ing</sub>) rather than dermal contact (EDI<sub>der</sub>). The metal species in each aquifer show that As has the lowest intake compared to other metals across all exposure routes and population groups. In contrast, the highest metal intake via ingestion in both children and adults in the Bangkok, Phra Pradaeng, and Nonthaburi aquifers is Ni. The highest metal intake via dermal contact in the Bangkok and Nonthaburi aquifers is Pb, while in the

Phra Pradaeng and Nakhon Luang aquifers, it is Zn. Key factors contributing to the varying levels of metal intake include the differing concentrations of heavy metals, leading to diverse intake levels, the higher ET value in children compared to adults, which results in greater metal intake in children, [46] and the higher absorption rate of the gastrointestinal system relative to dermal contact [58,59], making ingestion a more prominent route of exposure. Additionally, other factors influencing the differential intake include the skin barrier [60], which prevents absorption through the skin, and the distinct Kp values for each metal.

**Table 4.** Estimated daily intake (mg kg<sup>-1</sup> day<sup>-1</sup>) of heavy metals from ingestion and dermal exposure in children and adults across different aquifers.

Aquifer	Receptor	Exposure pathway	Average daily intake (mg kg <sup>-1</sup> day <sup>-1</sup> )			
			As	Ni	Pb	Zn
Bangkok	Children	Ingestion	0	1.711×10 <sup>-1</sup>	1.040×10 <sup>-2</sup>	7.560×10 <sup>-3</sup>
		Dermal	0	1.255×10 <sup>-4</sup>	1.525×10 <sup>-4</sup>	1.663×10 <sup>-5</sup>
	Adult	Ingestion	0	4.481×10 <sup>-2</sup>	2.724×10 <sup>-3</sup>	1.980×10 <sup>-3</sup>
		Dermal	0	4.399×10 <sup>-5</sup>	5.349×10 <sup>-5</sup>	5.832×10 <sup>-6</sup>
Phra Pradaeng	Children	Ingestion	1.200×10 <sup>-4</sup>	1.795×10 <sup>-1</sup>	8.960×10 <sup>-3</sup>	1.350×10 <sup>-1</sup>
		Dermal	4.400×10 <sup>-7</sup>	1.316×10 <sup>-4</sup>	1.314×10 <sup>-4</sup>	2.969×10 <sup>-4</sup>
	Adult	Ingestion	3.143×10 <sup>-5</sup>	4.702×10 <sup>-2</sup>	2.347×10 <sup>-3</sup>	3.535×10 <sup>-2</sup>
		Dermal	1.543×10 <sup>-7</sup>	4.615×10 <sup>-5</sup>	4.608×10 <sup>-5</sup>	1.041×10 <sup>-4</sup>
Nakhon Luang	Children	Ingestion	1.680×10 <sup>-3</sup>	1.535×10 <sup>-1</sup>	9.080×10 <sup>-3</sup>	1.717×10 <sup>-1</sup>
		Dermal	6.160×10 <sup>-6</sup>	1.126×10 <sup>-4</sup>	1.332×10 <sup>-4</sup>	3.778×10 <sup>-4</sup>
	Adult	Ingestion	4.400×10 <sup>-4</sup>	4.020×10 <sup>-2</sup>	2.378×10 <sup>-3</sup>	4.497×10 <sup>-2</sup>
		Dermal	2.160×10 <sup>-6</sup>	3.947×10 <sup>-5</sup>	4.670×10 <sup>-5</sup>	1.325×10 <sup>-4</sup>
Nonthaburi	Children	Ingestion	1.740×10 <sup>-3</sup>	1.172×10 <sup>-1</sup>	6.570×10 <sup>-3</sup>	3.129×10 <sup>-2</sup>
		Dermal	6.380×10 <sup>-6</sup>	8.591×10 <sup>-5</sup>	9.636×10 <sup>-5</sup>	6.884×10 <sup>-5</sup>
	Adult	Ingestion	4.557×10 <sup>-4</sup>	3.068×10 <sup>-2</sup>	1.721×10 <sup>-3</sup>	8.195×10 <sup>-3</sup>
		Dermal	2.237×10 <sup>-6</sup>	3.012×10 <sup>-5</sup>	3.379×10 <sup>-5</sup>	2.414×10 <sup>-5</sup>

Analysis of the trends in metal intake reveals that the trend for heavy metal intake via ingestion in the Bangkok aquifer for both children and adults is Ni > Pb > Zn > As, while via dermal contact, it follows the order Pb > Ni > Zn > As. The trend of dermal intake in the Bangkok aquifer is similar to that in Nonthaburi. In Nonthaburi, the trend of metal intake through ingestion mirrors that of Phra Pradaeng, with Ni > Zn > Pb > As, while the trend of ingestion through the gastrointestinal route in the

Nakhon Luang aquifer is Zn > Ni > Pb > As. The trend of dermal intake in the Phra Pradaeng aquifer is identical to that of ingestion in the Nakhon Luang aquifer, whereas the trend of dermal intake in the Nakhon Luang aquifer is Zn > Pb > Ni > As.

#### **Non-carcinogenic health risk assessment**

The non-carcinogenic health risk assessment from the calculation of HQ and HI is shown in Table 5.

**Table 5.** The hazard quotient (HQ) and hazard index (HI) values of heavy metals from ingestion and dermal exposure in children and adults across different aquifers.

Aquifer	Receptor	Exposure pathway	Hazard quotient				Hazard index
			As	Ni	Pb	Zn	
Bangkok	Children	Ingestion	0	8.554	7.429	$2.520 \times 10^{-2}$	16.322
		Dermal	0	$2.323 \times 10^{-2}$	$2.911 \times 10^{-1}$	$2.772 \times 10^{-4}$	
	Adult	Ingestion	0	2.240	1.946	$6.600 \times 10^{-3}$	
		Dermal	0	$8.147 \times 10^{-3}$	$1.021 \times 10^{-1}$	$9.720 \times 10^{-5}$	
Phra Pradaeng	Children	Ingestion	$4.000 \times 10^{-1}$	8.974	6.400	$4.499 \times 10^{-1}$	16.508
		Dermal	$3.577 \times 10^{-3}$	$2.437 \times 10^{-2}$	$2.508 \times 10^{-1}$	$4.949 \times 10^{-3}$	
	Adult	Ingestion	$1.048 \times 10^{-1}$	2.350	1.676	$1.178 \times 10^{-1}$	
		Dermal	$1.254 \times 10^{-3}$	$8.547 \times 10^{-3}$	$8.794 \times 10^{-2}$	$1.735 \times 10^{-3}$	
Nakhon Luang	Children	Ingestion	5.600	7.675	6.486	$5.724 \times 10^{-1}$	20.664
		Dermal	$5.008 \times 10^{-2}$	$2.085 \times 10^{-2}$	$2.541 \times 10^{-1}$	$6.296 \times 10^{-3}$	
	Adult	Ingestion	1.467	2.010	1.699	$1.499 \times 10^{-1}$	
		Dermal	$1.756 \times 10^{-2}$	$7.310 \times 10^{-3}$	$8.912 \times 10^{-2}$	$2.208 \times 10^{-3}$	
Nonthaburi	Children	Ingestion	5.800	5.858	4.693	$1.043 \times 10^{-1}$	16.707
		Dermal	$5.187 \times 10^{-2}$	$1.591 \times 10^{-2}$	$1.839 \times 10^{-1}$	$1.147 \times 10^{-3}$	
	Adult	Ingestion	1.519	1.534	1.229	$2.732 \times 10^{-2}$	
		Dermal	$1.819 \times 10^{-2}$	$5.579 \times 10^{-3}$	$6.448 \times 10^{-2}$	$4.023 \times 10^{-4}$	

The assessment of the health impacts and risks from exposure to As, Ni, Pb, and Zn via a single exposure pathway reveals that the HQ values for As, Ni, Pb, and Zn via ingestion in children range from 0 to 5.800, 5.858 to 8.974, 4.693 to 7.429, and  $2.520 \times 10^{-2}$  to  $5.725 \times 10^{-1}$ , respectively, while in adults, the ranges are 0 to 1.519, 1.534 to 2.350, 1.229 to 1.946, and  $6.600 \times 10^{-3}$  to  $1.499 \times 10^{-1}$ , respectively. The trend in HQ values shows that children have higher HQ values than adults, indicating that children are at a higher health risk from heavy metals than adults.

The assessment of health risks from exposure to heavy metals via ingestion shows that the only metal with no health risk through ingestion is Zn ( $HQ < 1$ ), while As, Ni, and Pb have unacceptable HQ values ( $HQ > 1$ ). Specifically, the HQ value for As in children from groundwater in the Nakhon Luang aquifer (5.600) and Nonthaburi aquifer (5.800), and in adults from groundwater in Nakhon Luang aquifer (1.467) and Nonthaburi aquifer (1.519) are all higher than 1. Additionally, the HQ values for Ni and Pb in both children and adults across all aquifers indicate significant health risks. In particular, the HQ values for Ni in children are 8.554, 8.974, 7.675, and 5.858 for the

Bangkok, Phra Pradaeng, Nakhon Luang, and Nonthaburi aquifers, respectively. and in adults, the values are 2.240, 2.350, 2.010, and 1.534. Pb HQ values for children from groundwater in Bangkok, Phra Pradaeng, Nakhon Luang, and Nonthaburi aquifer are 7.429, 6.400, 6.486, and 4.693, and in adults, the values are 1.946, 1.674, 1.699, and 1.229, respectively. This indicates that in the Bangkok and Phra Pradaeng aquifers, the concentrations of Ni and Pb pose a health risk from ingestion.

The health risk assessment from exposure to As, Ni, Pb, and Zn via contact shows that there is no health risk from exposure to these metals through contact for either children or adults ( $HQ < 1$ ). The metal that poses the least risk for both children and adults is As, with an HQ value of 0. The highest risk metal is Pb in children ( $2.911 \times 10^{-1}$ ).

Furthermore, the study also shows that the EDI is not the only factor contributing to higher health risks. Another important factor is the RefD for each metal, which represents the amount of a metal that can be safely consumed over a lifetime without adverse health effects. A lower RefD results in a higher HQ when compared to cases where the same amount of metal is ingested. The

analysis shows that Zn has the lowest RefD, which leads to lower HQ values for Zn, even though the intake of Zn via ingestion in the Nakhon Luang aquifer and via contact in the Phra Pradaeng and Nakhon Luang aquifer is high.

The health risk assessment from multiple exposure pathways shows that health risks are unacceptable ( $HI > 1$ ), with children's risk being 4 times higher than that of adults in all aquifers. The aquifer with the highest health risk is the Nakhon Luang aquifer, with an HI of 20.664 in children and 5.442 in adults. Other aquifers show similar trend in HI values for both children and adults as in the Nakhon Luang aquifer.

Based on the analysis, it can be concluded that diseases that could arise from health risks exceeding the acceptable level may result from exposure to As, Ni, and Pb with HQ values  $> 1$ . The potential health impacts include: 1) Exposure to As could affect the digestive

system, skin, nervous system, brain, and circulatory system [61]. 2) Exposure to Ni, although typically inhaled through industrial activities, can also pose risks when ingested from contaminated groundwater, leading to respiratory issues, heart disease, and, when in contact with the skin, inflammation [62]. Additionally, ingestion of Ni could affect the urinary system, especially the kidneys [63]. 3) Pb exposure in children poses a greater risk to the nervous system and brain than in adults, potentially lowering IQ and impairing memory. In adults, Pb exposure can lead to anemia, kidney disease, and disorders in the reproductive system, nervous system, and bones [64].

### Carcinogenic health risk assessment

The health risk assessment for carcinogenic heavy metals, including As, Ni, and Pb, was calculated using CR and TCR values, as shown in Table 6.

**Table 6.** The carcinogenic risk (CR) and total carcinogenic risk (TCR) values of heavy metals from ingestion and dermal exposure in children and adults across different aquifers.

Aquifer	Receptor	Exposure pathway	Carcinogenic risk				Total carcinogenic risk
			As	Ni	Pb	Zn	
Bangkok	Children	Ingestion	0	$2.910 \times 10^{-1}$	$8.840 \times 10^{-5}$	NC	$2.963 \times 10^{-1}$
		Dermal	0	$5.332 \times 10^{-3}$	NC	NC	
	Adult	Ingestion	0	$7.617 \times 10^{-2}$	$2.315 \times 10^{-5}$	NC	$7.806 \times 10^{-2}$
		Dermal	0	$1.870 \times 10^{-3}$	NC	NC	
Phra Pradaeng	Children	Ingestion	$1.800 \times 10^{-4}$	$3.051 \times 10^{-1}$	$7.616 \times 10^{-5}$	NC	$3.110 \times 10^{-1}$
		Dermal	$1.610 \times 10^{-6}$	$5.594 \times 10^{-3}$	NC	NC	
	Adult	Ingestion	$4.714 \times 10^{-5}$	$7.991 \times 10^{-2}$	$1.995 \times 10^{-5}$	NC	$8.194 \times 10^{-2}$
		Dermal	$5.647 \times 10^{-7}$	$1.961 \times 10^{-3}$	NC	NC	
Nakhon Luang	Children	Ingestion	$2.520 \times 10^{-3}$	$2.610 \times 10^{-1}$	$7.718 \times 10^{-5}$	NC	$2.684 \times 10^{-1}$
		Dermal	$2.255 \times 10^{-5}$	$4.784 \times 10^{-3}$	NC	NC	
	Adult	Ingestion	$6.600 \times 10^{-4}$	$6.834 \times 10^{-2}$	$2.021 \times 10^{-5}$	NC	$7.071 \times 10^{-2}$
		Dermal	$7.906 \times 10^{-6}$	$1.678 \times 10^{-3}$	NC	NC	
Nonthaburi	Children	Ingestion	$2.610 \times 10^{-3}$	$1.992 \times 10^{-1}$	$5.585 \times 10^{-5}$	NC	$2.055 \times 10^{-1}$
		Dermal	$2.335 \times 10^{-5}$	$3.651 \times 10^{-3}$	NC	NC	
	Adult	Ingestion	$6.836 \times 10^{-4}$	$5.216 \times 10^{-2}$	$1.463 \times 10^{-5}$	NC	$5.415 \times 10^{-2}$
		Dermal	$8.188 \times 10^{-6}$	$1.280 \times 10^{-3}$	NC	NC	

NC, not calculated

The CR values for single-element exposure to As, Ni, and Pb via ingestion in children ranged from 0 to  $2.601 \times 10^{-3}$ ,  $1.992 \times 10^{-1}$  to  $3.051 \times 10^{-1}$ , and  $5.585 \times 10^{-5}$  to  $8.840 \times 10^{-5}$ , respectively. In adults, the CR values ranged from 0 to  $6.836 \times 10^{-4}$ ,  $5.216 \times 10^{-2}$  to  $7.991 \times 10^{-2}$ ,

and  $1.463 \times 10^{-5}$  to  $2.315 \times 10^{-5}$ , respectively. For dermal exposure, the CR values for As and Ni in children were 0 to  $2.335 \times 10^{-5}$  and  $3.651 \times 10^{-3}$  to  $5.594 \times 10^{-3}$ , respectively. In adults, these values ranged from 0 to

$8.188 \times 10^{-6}$  and  $1.280 \times 10^{-3}$  to  $1.961 \times 10^{-3}$ , respectively.

The CR results indicate that the intake of Ni through ingestion and dermal exposure in both children and adults across all aquifers, as well as the intake of As through ingestion in children across all aquifers (except the Bangkok aquifer) and in adults in the Nakhon Luang and Nonthaburi aquifers, poses unacceptable health risks ( $CR > 10^{-4}$ ). In contrast, Pb intake via ingestion in both children and adults across all aquifers, As intake via dermal exposure in children in the Phra Pradaeng, Nakhon Luang, and Nonthaburi aquifers, and in adults in the Nakhon Luang and Nonthaburi aquifers, as well as As intake via ingestion in the Phra Pradaeng aquifer, fall within acceptable risk levels ( $10^{-6} < CR < 10^{-4}$ ). On the other hand, As intake through dermal exposure in adults in the Phra Pradaeng aquifer and ingestion and dermal exposure in all population groups in the Bangkok aquifer present no significant health risks. Therefore, Ni represents the most significant health risk in all population groups and aquifers, while As presents a risk in the Phra Pradaeng, Nakhon Luang, and Nonthaburi aquifers for specific exposure pathways and population groups.

The results of the multi-element exposure assessment using multiple exposure pathways based on TCR values show that health risks for both children and adults in the aquifers exceed acceptable levels ( $TCR > 10^{-4}$ ). The highest risks are found in the Phra Pradaeng aquifer, with TCR values of  $3.110 \times 10^{-1}$  for children and  $8.194 \times 10^{-2}$  for adults. The Nonthaburi aquifer also shows high risks, with TCR values of  $2.055 \times 10^{-1}$  for children and  $5.415 \times 10^{-2}$  for adults. Therefore, groundwater in these industrial areas should undergo quality improvement before use, as prolonged exposure could lead to cancer. Lead exposure could result in stomach, brain, and kidney cancers [65], while arsenic exposure could lead to stomach, skin, and lung cancers [66], and nickel exposure could cause lung and nasal cancers [67]. Methods for improving groundwater quality before use include chemical treatments, combined chemical and physical treatments, and biosorption techniques [68]. Additionally, development plans [69] and legislation to control and prevent heavy metal leakage from industrial processes into aquifers should be implemented to reduce health

risks associated with groundwater usage. In comparing the health risks from heavy metals in groundwater in this study with the risks from previous research conducted in the study area, this comparison revealed limitations due to the lack of data on health risk assessments from previous studies, particularly regarding contamination in deep aquifers. Sampling from these deep groundwater aquifers requires observation wells that reach significantly below the surface, and the number of such wells is limited, being under strict management by the responsible agencies. Furthermore, sampling from deep groundwater aquifers necessitates specialized tools and equipment, posing additional challenges in obtaining a sufficient amount of data.

## CONCLUSIONS

The groundwater in the studied industrial area is contaminated with As, Ni, Pb, and Zn, which are heavy metals commonly used in local industries. The assessment of contamination levels reveals that Zn is the only metal that does not exceed the standard in all groundwater aquifers. However, the average concentration of As in the Nakhon Luang, and Nonthaburi aquifers exceeds the USEPA and WHO standards. The average concentration of Pb and Ni also exceed the standards when compared to the USEPA, WHO, and NOAA groundwater standards. When groundwater from these areas is used, heavy metals can enter the body. The highest metal intake in each aquifer occurs through ingestion in children. Ni is the metal with the highest intake in the Bangkok, Phra Pradaeng, and Nonthaburi aquifers, while Zn is the metal with the highest intake in the Nakhon Luang aquifer.

Health risk assessment for non-carcinogenic effects from single-element exposure pathways shows that the health risk from Pb and Ni through ingestion in both children and adults exceeds the acceptable threshold. Meanwhile, As poses a non-carcinogenic risk only for children and adults through ingestion in the Nakhon Luang and Nonthaburi aquifers. Zn is the only metal that remains within acceptable risk levels in all aquifers.

Health risk assessment from multi-element exposure pathways shows that the risk from heavy metals in both children and adults across all aquifers exceeds the

acceptable threshold, which could lead to non-cancer diseases from groundwater utilization.

Regarding carcinogenic health risks from single-element exposure pathways, Ni poses an unacceptable health risk through ingestion and dermal contact across all aquifers for all population groups. As exposure through dermal contact and ingestion in both children and adults in the Bangkok aquifer, as well as through dermal contact in adults in the Phra Pradaeng aquifer, shows no significant risk, while other exposure pathways and population groups for As and ingestion in all population groups for Pb pose a risk. Based on the findings, it can be concluded that groundwater used in areas with heavy metal contamination will pose health risks. Therefore, groundwater should be treated before use to mitigate potential health risks from heavy metal exposure.

#### ACKNOWLEDGEMENTS

The authors would like to express their sincere gratitude to the Faculty of Science and Technology, Valaya Alongkorn Rajabhat University under the Royal Patronage, for their essential support, including providing laboratory equipment and chemicals for this research. We also appreciate the Department of Groundwater Resources for their assistance during the field sampling process. Finally, we extend our thanks to all individuals involved in the groundwater sampling for their invaluable contributions to this study.

#### Conflict of interest

The authors declares no conflict of interest.

#### REFERENCES

- Lubal M.J., 2024. Impact of Heavy Metal Pollution on the Environment. *Uttar Pradesh Journal of Zoology*. 45(11), 97–105. doi: 10.56557/upjoz/2024/v45i114074.
- Zaynab M., Al-Yahyai R., Ameen A., Sharif Y., Ali L., Fatima M., Khan K.A., Li S., 2022. Health and environmental effects of heavy metals. *Journal of King Saud University - Science*. 34(1), 101653. doi: 10.1016/j.jksus.2021.101653.
- Ullah Z., Rashid A., Ghani J., Nawab J., Zeng X.C., Shah M., Alrefaei A.F., Kamel M., Aleya L., Abdel-Daim M.M., Iqbal J., 2022. Groundwater contamination through potentially harmful metals. *Frontiers in environmental science*. 2022(10), 1054924. doi: 10.3389/fpls.2022.1054924.
- Gupta M., Dwivedi V., Kumar S., Patel A., Niazi P., Yadav V.K., 2024. Lead toxicity in plants: mechanistic insights into toxicity, physiological responses of plants and mitigation strategies. *Plant Signaling & Behavior*. 19(1), 2365576. doi: 10.1080/15592324.2024.2365576.
- Jomova M., Makova M., Alomar S.Y., Alwasel S.H., Nepovimova E., Kuca K., Rhodes C.J., Valko M., 2022. Essential metals in health and disease. *Chemo-Biological Interactions*. 2022(367), 110173. doi: 10.1016/j.cbi.2022.110173.
- Choi S., Hong D.K., Choi B.Y., Suh S.W., 2020. Zinc in the Brain: Friend or Foe?. *Int J Mol Sci*. 21(23), 8941. doi: 10.3390/ijms21238941.
- Timothy N., Williams E.T., 2019. Environmental Pollution by Heavy Metal: An Overview. *International Journal of Environmental Chemistry*. 3(2), 72-82. doi: 10.11648/j.ijec.20190302.14.
- Tangkan S., Sirisetpop, 2025. Health risk assessment of heavy metals in Cassava cultivated on leachate-contaminated soil during the early transition from landfilling to co-landfilling with incineration. *EQM – International Journal of Environmental Quality*. 2025(68), 63-75. doi: 10.6092/issn.2281-4485/21367.
- Department of Industrial Works. Summary of statistics on the number of industrial plants with a Type 3 operating license and those with a Type 2 notification in Bangkok and regional areas in 2022. <https://www.diw.go.th/webdiw/wp-content/uploads/2024/10/facyear2565.xlsx> (Accessed Dec 2, 2024).
- Zhang W., Xin C., Yu S., 2023. A Review of Heavy Metal Migration and Its Influencing Factors in Karst Groundwater, Northern and Southern China. *Water*. 15(20), 3690. doi: 10.3390/w15203690.
- Ogbaran N., Uguru H., 2021. Assessment of Groundwater Quality Around an Active Dumpsite using Pollution Index. *Civil Eng Res J*. 11(3), 555814. doi: 10.19080/CERJ.2021.11.555814
- Fang H., Wang X., Xia D., Zhu J., Yu W., Su Y., Zeng J., Zhang Y., Lin X., Lei Y., Qiu J., 2022. Improvement of Ecological Risk Considering Heavy Metal in Soil and Groundwater Surrounding

- Electroplating Factories. Processes. 10(7), 1267. doi: 10.3390/pr10071267.
13. Veskovi J., Onjia A., 2024. Environmental Implications of the Soil-to-Groundwater Migration of Heavy Metals in Mining Area Hotspots. Metals. 14(6), 719. doi: 10.3390/met14060719.
14. Department of Disease Control. 2022. A Lesson Learned Report: Surveillance of Health Impacts from Environmental Pollution: A Case Study of the Explosion and Fire at Ming Dih Chemical Company in Samut Prakan Province. Ministry of Public Health.
15. Department of Industrial Works. List of industrial plants in Bang Pu Industrial Estate. <https://bit.ly/4geMwKP> (Accessed Dec 6, 2024).
16. Pavithra N., Ramakrishnaiah C R., 2024. Heavy Metal Analysis and Health Risk Assessment of Groundwater and Soil in and Around Peenya Industrial Area, Bengaluru. Ecol Eng Environ Technol. 25(7), 63-79. doi: 10.12912/27197050/188008.
17. Daud M.K., Nafees M., Ali S., Rizwan M., Bajwa R.A., Shakoor M.B., Arshad M.U., Chatha S.A.S., Deeba F., Murad W., Malook I., Shui Jin Zhu S.J., 2017. Drinking water quality status and contamination in Pakistan. BioMed Res Int. 2017(3), 1–18. doi: 10.1155/2017/7908183.
18. Dippong T., Hoaghia M.A., Senila M., 2022. Appraisal of heavy metal pollution in alluvial aquifers. Study case on the protected area of Ronișoara Forest, Romania. Ecological Indicators. 2022(143), 109347. doi: 10.1016/j.ecolind.2022.109347.
19. Caporale A.G., Violante A., 2015. Chemical Processes Affecting the Mobility of Heavy Metals and Metalloids in Soil Environments. Current Pollution Reports. 2015(2), 15–27. doi: 10.1007/s40726-015-0024-y.
20. Mali M., Alfio M.R., Balacco G., Ranieri G., Specchio V., Fidelibus M.D., 2024. Mobility of trace elements in a coastal contaminated site under groundwater salinization dynamics. Scientific Reports. 2024(14), 24859. doi: 10.1038/s41598-024-75974-1.
21. Department of Groundwater Resources. Groundwater Situation in Bangkok and perimeter in 2012. <https://bit.ly/4ay7tiG> (Accessed Dec 7, 2024).
22. Rodger B.B., Andrew E.D., Eugene W.R., Standard methods for the examination of water and wastewater, 23rd ed., American public health association: Washington, 2017.
23. USEPA. Method 200.7: Determination of Metals and Trace Elements in Water and Wastes by Inductively Coupled Plasma-Atomic Emission Spectrometry, Revision 4.4; U.S. Environmental Protection Agency: Cincinnati, OH, 1994.
24. Wilkin R.T., Lee T.R., Beak D.G., Anderson R., Burns B., 2018. Groundwater co-contaminant behavior of arsenic and selenium at a lead and zinc smelting facility. Applied Geochemistry. 2018(89), 255–264. doi: 10.1016/j.apgeochem.2017.12.011.
25. USEPA. Method 6010D (SW-846): Inductively Coupled Plasma-Atomic Emission Spectrometry, Revision 4; U.S. Environmental Protection Agency: Washington, D.C., 2014.
26. USEPA. Risk Assessment Guidance for Superfund. Human Health Evaluation Manual (Part A), Vol. 1; U.S. Environmental Protection Agency: Washington, D.C., 1989.
27. Myers R.A., Gyimah E., Gbemadu K., Osei B., Akoto O., 2023. Appraising groundwater quality and the associated health risks of heavy metal contamination at Suame magazine. Scientific African. 21, e01794. doi: 10.1016/j.sciaf.2023.e01794.
28. Edokpayi J.N., Enitan A.M., Mutleni N., Odiyo J.O., 2018. Evaluation of water quality and human risk assessment due to heavy metals in groundwater around Muledane area of Vhembe District, Limpopo Province, South Africa. Edokpayi et al. Chemistry Central Journal. 2018(2), 1-16. doi: 10.1186/s13065-017-0369-y.
29. USEPA. Exposure factors handbook 2011 edition (final report). <https://cfpub.epa.gov/ncea/risk/recordisplay.cfm?deid=236252> (Accessed Dec 10, 2024).
30. Guo Y., Huang M., You W., Luxiang Cai L., Yong Hong Y., Xiao Q., Zheng X., Lin R., 2020. Spatial analysis and risk assessment of heavy metal pollution in rice in Fujian Province, China. Frontiers in Environmental Science. 2020(10), 82340. doi: 10.3389/fenvs.2022.1082340.
31. Akoto O., Samuel A., Gladys L., Sarah O.A.A., Apau J., Opoku F., 2022. Assessment of groundwater quality from some hostels around Kwame Nkrumah University

- of Science and Technology. *Scientific African*. 2022(17), e01361. doi: 10.1016/j.sciaf.2022.e01361.
32. Adimalla N., 2019. Heavy metals contamination in urban surface soils of Medak province, India, and its risk assessment and spatial distribution. *Environ Geochem Health*. 2020(42), 59–75.
33. USEPA. Integrated Risk Information System (IRIS); U.S. Environmental Protection Agency: Washington, D.C., USA, 2010.
34. Yin S., Feng C., Li Y., Yin L., Shen Z., 2015. Heavy metal pollution in the surface water of the Yangtze Estuary: a 5-year follow-up study. *Chemosphere*. 2015(138), 718–725. doi: 10.1016/j.chemosphere.2015.07.060.
35. Hasan M.F., Nur-E-Alam M., Salam M.A., Rahman H., Paul S.C., Rak A.E., Ambade B., Towfiqul Islam A.R.M., 2021. Health Risk and Water Quality Assessment of Surface Water in an Urban River of Bangladesh. *Sustainability*. 13(12), 6832. doi: 10.3390/su13126832.
36. Wang K., Aji D., Li P., Hu C., 2024. Characterization of heavy metal contamination in wetland sediments of Bosten lake and evaluation of potential ecological risk. *Frontiers in Environmental Science*. 2024(12), 1398849.
37. Mohanty S., Nayak R.K., Jena B., Padhan K., Mohapatra K.K., Sahoo S.K., Dash P.K., Das J., Behera S.K., Sahu A., Nayak J.K., Padhan S., Datta D., 2023. Heavy metal contamination in rice, pulses, and vegetables from CKDu-endemic areas in Cuttack district, India: a health risk assessment. *Front. Environ. Sci*. 2023(11), 1248373. doi: :10.3389/fenvs.2023.1248373.
38. Haque R.M., Ali M.M., Ahmed W., Rahman M.M., 2022. Assessment of metal(loid)s pollution in water and sediment from an urban river in Bangladesh: An ecological and health risk appraisals. *Case Studies in Chemical and Environmental Engineering*. 2022(6), 100272.
39. Abdullah N.H., Kean O.B., Hirmizi M.N., Yusoff N., Sabarudin A.R.N.M., 2020. Method validation of heavy metals determination in traditional herbal tablet, capsule and liquid by graphite furnace atomic absorption spectrometer and flow injection for atomic spectroscopy hydride system. *Asian Journal Of Pharmacognosy*. 4(3), 37-45.
40. Mwakisunga B., Pratap H.B., Machiwa J.F., Stephano F., 2021. Heavy Metal Contamination and Potential Ecological Risks in Surface Sediments along Dar es Salaam Harbour Channel. *Tanzania Journal of Science*. 47(5), 1606-1621. doi: 10.1016/j.marpolbul.2016.09.038. doi:10.4314/tjs.v47i5.11.
41. AOAC international. AOAC Guidelines for Single Laboratory Validation of Chemical Methods for Dietary Supplements and Botanicals. [https://s27415.pcdn.co/wp-content/uploads/2020/01/64ER20-7/Validation\\_Methods/d-AOAC\\_Guidelines\\_For\\_Single\\_Laboratory\\_Validation\\_Dietary\\_Supplements\\_and\\_Botanicals.pdf](https://s27415.pcdn.co/wp-content/uploads/2020/01/64ER20-7/Validation_Methods/d-AOAC_Guidelines_For_Single_Laboratory_Validation_Dietary_Supplements_and_Botanicals.pdf) (Accessed Dec 8, 2024).
42. Khodami S., Surif M., Omar W.M.W., Daryanabard R., 2017. Assessment of heavy metal pollution in surface sediments of the Bayan Lepas area, Penang, Malaysia. *Marine Pollution Bulletin*. 2017(1), 615–622. doi: 10.1016/j.marpolbul.2016.09.038.
43. Spencer K.L., MacLeod C.L., 2002. Distribution and partitioning of heavy metals in estuarine sediment cores and implications for the use of sediment quality standards. *Hydrology and Earth System Sciences*. 6(6), 989-998. doi : 10.5194/hess-6-989-2002.
44. Kubier A., Hamer K., Pichler T., 2019. Cadmium Background Levels in Groundwater in an Area Dominated by Agriculture. *Integrated Environmental Assessment and Management*. 16(1), 103-113. doi: 10.1002/ieam.4198.
45. Lin M., Gui H., Peng W., Chen S., 2014. Heavy Metals Characteristics in Deep Groundwater of Coal Mining Area, Northern Anhui Province. *An Interdisciplinary Response to Mine Water Challenges*. 9(319), 171-177.
46. USEPA. National Primary Drinking Water Regulations. <https://www.epa.gov/ground-water-and-drinking-water/national-primary-drinking-water-regulations> (Accessed Dec 12, 2024).
47. USEPA. Technical Factsheet on: NICKEL. <https://archive.epa.gov/water/archive/web/pdf/archived-technical-fact-sheet-on-nickel.pdf> (Accessed Dec 13, 2024).
48. Tangkan S., Phanthasa S., 2024. Health Risk and Contamination Assessment of Heavy Metals in

Groundwater around Municipal Landfill, Ayutthaya Province. *Huachiew Chalermprakiet Science and Technology Journal*. 10(2), 82-97.

49. National Oceanic and Atmospheric Administration. screening quick reference tables. <https://www.nrc.gov/docs/ML0720/ML072040354.pdf> (Accessed Dec 13, 2024).

50. Eziz M., Sidikjan N., Zhong Q., Rixit A., Li X., 2023. Distribution, pollution levels, and health risk assessment of heavy metals in groundwater in the main pepper production area of China. *De Gruyter*. 2023(15), 20220491. doi: 10.1515/geo-2022-0491.

51. Ishtiaq M., Khan M.J., Khan S.A., Ghani J., Ullah Z., Nawab J., Alrefaei A.F., Almutairi M.H., Alharbi S.N., 2024. Potentially harmful elements and health risk assessment in groundwater of urban industrial areas. *Frontiers in Environmental Science*. 2024(12), 1332965. doi: 10.3389/fenvs.2024.1332965.

52. Mikongoti S.B., Jeremiah J.M., Kaunga D.L., 2023. Investigation of Heavy Metal Composition and Associated Health Risks from Selected Groundwater Wells in Temeke, Dar-es-Salaam. *Tanzania Journal of Engineering and Technology*. 42(4), 58 - 74. doi: 10.52339/tjet.v42i4.849.

53. Perumal M., Velusamy S.G., Subramanian M., Velmurugan P.M., Raj T.N., Reddy M.S., 2022. Heavy metal contamination and the assessment of health risks in groundwater in Arani industrial zones in Southern India. *Arabian Journal of Geosciences*. 2022(15), 948. doi: 10.1007/s12517-022-10223-1.

54. Ullah Z., Rashid A., Ghani J., Nawab J., Zeng X.C., Shah M., Alrefaei A.F., Kamel A., Aleya L., Abdel-Daim M.M., Iqbal J., 2022. Groundwater contamination through potentially harmful metals and its implications in groundwater management. *Frontiers in Environmental Science*. 2022(10), 1021596. doi: 10.3389/fenvs.2022.1021596.

55. Rahman M.A.T.M.T., Paul M., Bhoumik N., Hassan M., Alam M.K., Aktar Z., 2020. Heavy metal pollution assessment in the groundwater of the Meghna Ghat industrial area, Bangladesh, by using water pollution indices approach. *Applied Water Science*. 2020(10), 186. doi: 10.1007/s13201-020-01266-4.

56. Neporozhniaia I., Snetkova I., 2021. Correlation analysis of heavy metals content in bottom sediments of

the shallow zone Sheksninskaya spur of the Rybinsk reservoir in the city of Cherepovets (Russia, Vologda oblast), *E3S Web of Conferences* 265. 2021(265), 02017. doi: 10.1051/e3sconf/202126502017.

57. Raja V., Lakshmi R.V., Sekar C.P., Chidambaram S., Neelakantan M.A., 2021. Health Risk Assessment of Heavy Metals in Groundwater of Industrial Township Virudhunagar, Tamil Nadu, India. *Archives of Environmental Contamination and Toxicology*. 2021(80), 144–163. doi: 10.1007/s00244-020-00795-y.

58. Qu L., Huang H., Xia F., Liu Y., Dahlgren R.A., Zhang M., Mei K., 2018. Risk analysis of heavy metal concentration in surface waters across the rural-urban interface of the Wen-Rui Tang River, China. *Environmental Pollution*. 2018(237), 639-649. doi: 10.1016/j.envpol.2018.02.020.

59. Government of Canada. Guidance on Heavy Metal Impurities in Cosmetics. <https://www.canada.ca/en/health-canada/services/consumer-product-safety/reports-publications/industry-professionals/guidance-heavy-metal-impurities-cosmetics.html> (Accessed Dec 14, 2024).

60. Chavatte L., Juan M., Mounicou S., Noblesse E.L., Pays K., Nizard C., Bulteau A.L., 2020. Elemental and molecular imaging of human full thickness skin after exposure to heavy metals. *12(10)*, 1555-1562.

61. Briffa J., Sinagra E., Blundell R., 2020. Heavy metal pollution in the environment and their toxicological effects on humans. *Heliyon*. 2020(6), e04691. doi: 10.1016/j.heliyon.2020.e04691.

62. Chen, Q. Y.; Brocato, J.; Laulich, F.; Costa, M. Mechanisms of Nickel Carcinogenesis. In *Essential and Non-Essential Metals. Molecular and Integrative Toxicology*; Mudipalli, A., Zelikoff, J. T., Eds.; Springer International Publishing AG: New York, 2017. pp. 181–197.

63. Genchi G., Carocci A., Lauria G., Sinicropi M.S., Catalano A., 2020. Nickel: Human Health and Environmental Toxicology. *International Journal of Environmental Research and Public Health*. 17(3), 679. doi: 10.3390/ijerph17030679.

64. World Health Organization. Exposure to lead: a major public health concern. <https://iris.who.int/bitstream/handle/10665/372293/>

9789240078130-eng.pdf?sequence=1 (Accessed Dec 16, 2024).

65. Liao L.M., Friesen M.C., Xiang Y.B., Cai H., Koh D.H., Ji B.T., Yang G., Li H.L., Locke S.J., Rothman N., Zheng W., Gao Y.T., Shu X.O., Purdue M.P., 2016.

Occupational Lead Exposure and Associations with Selected Cancers: The Shanghai Men's and Women's Health Study Cohorts. *Environmental Health Perspectives*. 2016(124), 97-103. doi: 10.1289/ehp.1408171.

66. Kasmi S., Moser L., Gonvers S., Dormond O., Demartines N., Labгаа I., 2023. Carcinogenic effect of arsenic in digestive cancers: a systematic review. *Environmental Health*. 2023(22), 36. doi: 10.1186/s12940-023-00988-7.

67. Canadian Environmental Protection Act. Nickel and its Compounds, 1st ed., National printers(Ottawa) inc.: Canada, 1994.

68. Hashim M.A., Mukhopadhyay S., Sahu J.N., Sengupta B., 2011. Remediation technologies for heavy metal contaminated groundwater. *Journal of Environmental Management*. 2011(92), 2355-2388. doi: 10.1016/j.jenvman.2011.06.009.

69. Andabili N.R., Safaripour M., 2021. Pollution Assessment of Trace Metals in Ground Waters (Case Study: Meshgin Shahr County). *Journal of Chemical Health Risks*. 11(4), 375-382. doi: 10.22034/jchr.2020.1902942.1150.

