



## ORIGINAL ARTICLE

## Trace Element Content and Potential Human Health Risk from Consumption of the Deep-water Rose Shrimp *Parapenaeus longirostris* (Crustacea: Decapoda) from Pagasitikos Gulf, Greece

Konstantinos Skordas\*, Alexios Lolas, Christina Gounari, Konstantinos Georgiou, Nikolaos Neofitou, Dimitrios Vafidis

Department of Ichthyology and Aquatic Environment, School of Agricultural Sciences, University of Thessaly, Fytoko Street, 38445, Magnesia, Greece

(Received: 6 March 2022

Accepted: 10 May 2021)

## KEYWORDS

Heavy metals;  
Target Hazard  
Quotients;  
Estimated Weekly  
Intakes;  
Aegean Sea;  
East Mediterranean

**ABSTRACT:** The content of As, Cd, Cr, Cu, Fe, Mn, Ni, Pb, and Zn as the whole-body burden and the content of Cd and Pb in muscle tissue of *Parapenaeus longirostris*, were evaluated in shrimp collected from the Pagasitikos Gulf, Greece. The whole-body burden of Fe was  $102 \pm 39$ , Mn  $71 \pm 23$ , As  $64 \pm 33$ , Zn  $62 \pm 8$ , Cu  $34 \pm 16$ , Ni  $4.6 \pm 2.9$ , Cr  $1.01 \pm 0.39$ , Cd  $0.93 \pm 0.33$  and Pb  $0.88 \pm 0.47$  ppm wet weight. Muscle tissue content for Cd was  $0.47 \pm 0.08$  and for Pb  $0.31 \pm 0.06$  ppm wet weight. The estimation of weekly intakes and target hazard quotients for the potentially toxic elements Cd and Pb revealed that rose shrimp from Pagasitikos Gulf could be considered safe for human consumption, probably with a general advisory to avoid the consumption of anything other than the muscle tissue.

## INTRODUCTION

In the marine environment, elements pose a serious threat due to their toxic and non-biodegradable and long persistent properties [1] and they may further degrade the ecosystem [2, 3], thus accumulating in marine organisms and subsequently transferred to humans through the food chain, sometimes, in concentrations which are potentially toxic [4]. That potential toxicity can be of serious consequence, both for the environment and human health. High bioavailability of trace elements can, on one hand, have a direct toxic effect on the survival and growth of marine organisms, thus affecting community structure and production, while on the other hand; contaminated seafood might cause health implications when consumed by humans [5, 6].

Marine crustaceans and especially shrimps are high-value target species with great commercial importance and are

widely consumed across the Mediterranean [7, 8]. In some cases, they tend to accumulate more trace elements than fish, as a result of differences in the evolutionary strategies adopted by the different phyla [9]. In the Eastern Mediterranean region, they constitute a major component of the so-called “Mediterranean diet” [10], thus, increasing the need to determine the content of trace elements in local populations and to assess the health risk their consumption poses.

The deepwater rose shrimp *Parapenaeus longirostris* is considered the most important species (in catchweight, landings, and value) among commercial crustaceans along with the coasts of Spain, France, Italy, and Greece [11-13]. Owing to its abundance and high economic importance, there are many studies regarding the aspects of ecology and biology of *P. longirostris* (e.g. [11, 14,

\*Corresponding author: kskord@uth.gr (K. Skordas)

DOI: 10.22034/JCHR.2022.1956070.1541

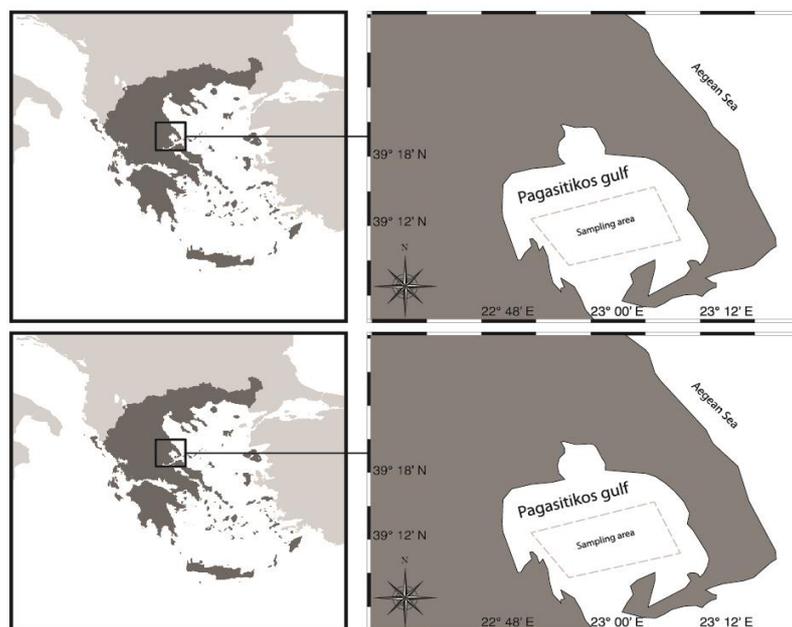
and 15] and references within), contrary to very few studies focusing on trace element content [10, 16-21]. The present study aimed to provide new information regarding the content of several trace elements (namely As, Cd, Cr, Cu, Fe, Mn, Ni, Pb, and Zn), in local deep-water rose shrimp stocks and, also, to assess the exposure to these potentially toxic elements from their consumption.

## MATERIALS AND METHODS

Many studies on trace element contamination usually focus on muscle tissue since it is the obvious edible part. In Greece, however, it is a very common cooking practice in local restaurants to serve fried rose shrimp which are consumed as whole shrimps. It was, thus, decided to estimate the content of the 9 trace elements in the whole body of *P. longirostris*, but also estimate the Cd and Pb content in the muscle tissue in particular, since these two elements can have high nephrotoxic effects [22] and they are generally considered a hazard for human health. In fact, along with Hg, they are the only trace elements with regulations regarding their maximum

concentration in seafood in the EU (EC Regulation No 1881/2006).

Rose shrimp samples were collected in June 2019, during an experimental trawl fishing expedition in the Pagasitikos Gulf (Figure 1). Shrimps below the minimum commercial size (carapace length < 20 mm), according to Greek fisheries regulation [12], were discarded and the rest of the catch was placed in polystyrene boxes, packed with ice flakes, and transported to the laboratory for further processing. Care was taken not to include shrimps with obvious signs of molting, to avoid misrepresentation of concentration levels of the elements, since crustacean exoskeleton may contain a significant amount of elements, but this may be reduced before molting [23]. In the laboratory, each rose shrimp was measured (Carapace Length, CL) to the nearest mm, for the estimation of size–frequency distribution of the sampled population. Also, they were weighed (Total Weight, TW) to the nearest 0.01 g, sexed, and finally, placed in individually labeled plastic bags and temporarily stored in a refrigerator (4°C), before sample preparation for the analytical procedures.



**Figure 1.** Map of the sampling area and the location of the Pagasitikos Gulf.

For the estimation of trace element content as whole-body burden 10 composite samples for each sex were prepared. Each sample contained 10 – 15 whole individuals of similar size, representing all the estimated CL size classes. For the estimation of Cd and Pb content

in the muscle tissue of *P. longirostris*, another set of 10 composite samples for each sex were created, only this time each rose shrimp was carefully dissected for the separation of muscle tissue from the carapace, and the visceral package (Figure 2). All samples were stored at -

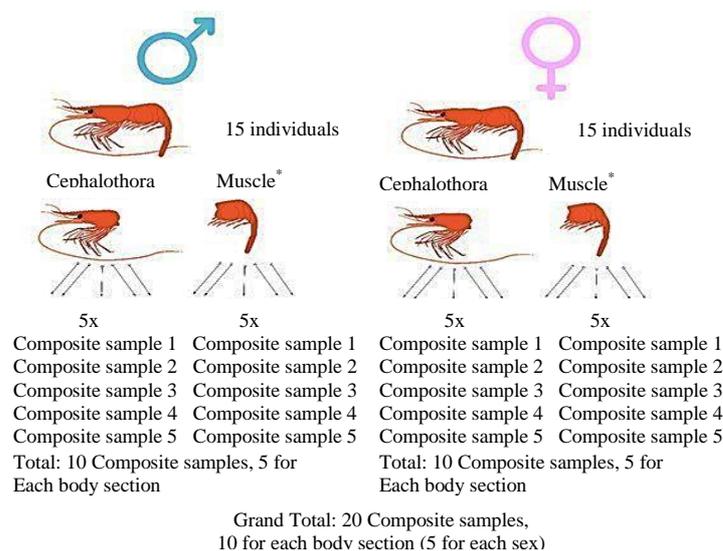
20°C, until further processed.

Before the detection analysis, the samples were removed from storage and thawed for 24 h at ambient room temperature and

Then, dried in an air-flow oven at 70°C, until constant weight. Each sample was homogenized and reduced to a fine powder, using mortar and pestle. Using a high precision analytical balance (MS104TS/A00, Mettler-Toledo)  $0.5 \pm 0.0001$  g from each sample were added in a Teflon vessel, along with 9 ml of HNO<sub>3</sub>, 2 ml H<sub>2</sub>O<sub>2</sub>, and 1 ml of ultra-pure water. The vessels were placed in a microwave wet-digestion system (Microwave 3000, Anton Paar) for 15 min, following the guidelines set by the EPA 3052 Method. After the conclusion of the digestion process, the vessels were cooled at ambient

room temperature and their contents were transferred into a volumetric flask, filled up to 100 ml with ultra-pure water, and stored at -4°C, until the detection analyses.

For the detection of Cu, Fe, Mn, and Zn the samples were analyzed by Flame Atomic Absorption Spectrometry (FAAS), using an atomic absorption spectrometer (AAnalyst 400, PerkinElmer). For the detection of As, Cd, Cr, Ni, and Pb, Graphite Furnace Atomic Absorption Spectrometry (GFAAS) was used (HGA 900 graphite furnace, equipped with Zeeman effect background correction and an AS 800 autosampler, by PerkinElmer). All the analyses were made in the Laboratory of Marine Biology, Department of Ichthyology and Aquatic Environment, School of Agricultural Sciences, University of Thessaly.



\*Muscle tissue was extracted from the abdomen by removing the exoskeleton and visceral package.

**Figure 2.** Composite sample creation process.

All samples were analyzed in triplicates and the concentration value for each element was represented by the median, rather than the mean of the results, because of their non-normal distribution [24]. The total content for each element was calculated by the WinLab32™ software (PerkinElmer) and expressed in ppm ( $\mu\text{g g}^{-1}$ ) wet weight.

All chemicals were of analytical reagent grade (Sigma-Aldrich). The ultra-pure water was double-distilled, using a Millipore Direct-Q-UV water purifier. All glassware was cleaned before use by soaking in 10% HNO<sub>3</sub> for 48 h and rinsing with Milli-Q water. Certified reference materials (NIST 1577b, NIST 2976, and BCR 61) were used as samples for method validation, with a satisfactory

recovery (92 – 103% for all measured elements).

The estimated weekly intakes (EWI) and target hazard quotient (THQ) of the trace elements Pb and Cd were calculated following the formulas described by Storelli [25] and Kalogeropoulos et al. [10], both for the whole body and the muscle tissue samples. The Provisional Tolerable Weekly Intakes (PTWI) of  $2.5 \mu\text{g kg}^{-1}$  body weight for Cd and  $25 \mu\text{g kg}^{-1}$  body weight for Pb were used [26]. The Total THQ (TTHQ), which is the sum of the respective THQs from Cd and Pb, was also calculated, as an indicator of the health risk from the consumption of *P. longirostris* [10, 25].

The null hypothesis of no significant differences in trace element content between the sexes was assessed using the

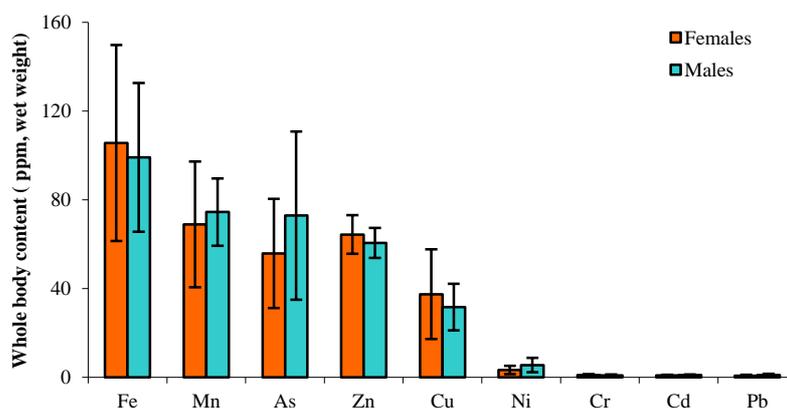
student's t-test. Spearman's correlation coefficient was used to measure the strength of the potential associations between elements. All the statistical analyses were performed by the STATGRAPHICS Centurion software package (v.16.1.11) and values of  $p < 0.05$  were considered significant.

## RESULTS AND DISCUSSION

The results from the estimation of the content of trace elements as a whole-body burden are presented in Table

**Table 1.** Mean concentration (ppm wet weight)  $\pm$  Standard Deviation and range of individual values of the 9 trace elements in composite samples from the whole body of *P. longirostris*.

| Element       | Fe           | Mn          | As          | Zn         | Cu          | Ni            | Cr              | Cd              | Pb              |
|---------------|--------------|-------------|-------------|------------|-------------|---------------|-----------------|-----------------|-----------------|
| Mean $\pm$ SD | 102 $\pm$ 39 | 71 $\pm$ 23 | 64 $\pm$ 33 | 62 $\pm$ 8 | 34 $\pm$ 16 | 4.6 $\pm$ 2.9 | 1.01 $\pm$ 0.39 | 0.93 $\pm$ 0.33 | 0.88 $\pm$ 0.47 |
| Range         | 49-201       | 23-138      | 26-171      | 46-80      | 14-85       | 0.62-12.1     | 0.62-2.22       | 0.61-1.88       | 0.21-2.11       |



**Figure 3.** Average content (ppm, wet weight) of trace elements in the body of female and male *P. longirostris* shrimps (Bars represent standard deviations).

According to Spearman's correlation analysis, Fe had weak, but significant positive relationships with Mn and Zn, while arsenic had weak, but significant positive relationships with Mn and Cu (Table 2). Usually, a positive correlation between elements can be regarded as indicative of particular biochemical pathways [23] or it could suggest that these elements share common sources, have mutual dependence, or have identical behavior during transport [27].

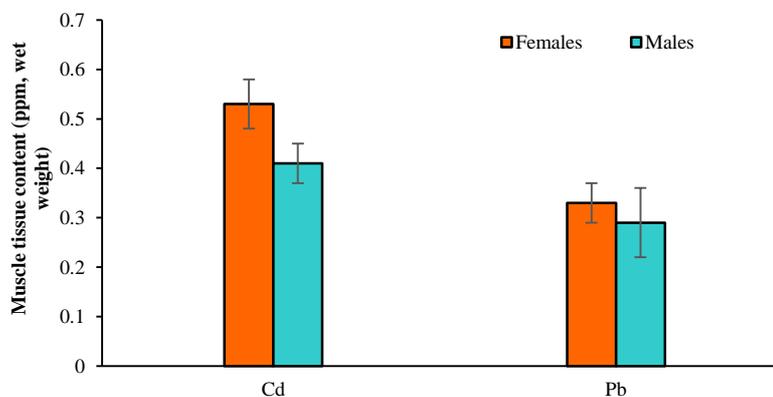
**Table 2.** Spearman ranks correlations between each pair of trace elements, based on whole-body content.

|    | Mn     | Zn     | Ni     | Cd     | Cu     | Pb     | As     | Cr    |
|----|--------|--------|--------|--------|--------|--------|--------|-------|
| Fe | *0.427 | *0.578 | -0.119 | -0.249 | 0.137  | -0.004 | 0.136  | 0.280 |
| Mn |        | -0.027 | 0.386  | -0.181 | 0.122  | -0.374 | *0.457 | 0.079 |
| Zn |        |        | -0.319 | -0.196 | 0.186  | 0.352  | -0.124 | 0.205 |
| Ni |        |        |        | -0.282 | 0.068  | -0.400 | 0.330  | 0.086 |
| Cd |        |        |        |        | -0.188 | -0.429 | -0.228 | 0.280 |
| Cu |        |        |        |        |        | -0.123 | *0.422 | 0.231 |
| Pb |        |        |        |        |        |        | 0.242  | 0.039 |
| As |        |        |        |        |        |        |        | 0.167 |

\*Indicates significant correlation,  $p < 0.05$

Regarding the content of Cd and Pb in the muscle tissue of *P. longirostris*, no differences were detected between the sexes ( $p > 0.05$ ) for either element. The average content of Cd in shrimp muscle tissue was  $0.47 \pm 0.08$  (0.36-0.61) and the respective content of Pb was

$0.31 \pm 0.06$  (0.20-0.39) ppm wet weight (Figure 4). The estimation results for all trace elements of the present study, along with the findings of similar studies, regarding *P. longirostris* and other shrimp species in the Eastern Mediterranean region, are presented in Table 3.



**Figure 4.** Average muscle tissue content (ppm, wet weight) of Cd and Pb in female and male *P. longirostris* shrimps (Bars represent standard deviations).

The estimated weekly intakes (EWI) and the target hazard quotients (THQ) for Cd and Pb, both from the whole body and muscle tissue of *P. longirostris* are shown in Table 4. Trace element intake through the consumption of whole individuals was estimated at  $0.59 \mu\text{g kg}^{-1}$  body weight for Cd and  $0.56 \mu\text{g kg}^{-1}$  body weight for Pb, representing 23.6% and 2.2% of the PTWI for each element. The respective intake from the consumption of muscle tissue was estimated at  $0.30 \mu\text{g kg}^{-1}$  body weight for Cd and  $0.20 \mu\text{g kg}^{-1}$  body weight for Pb, representing 11.9% and 0.8% of the PTWI for each element.

Many of the studied elements (Fe, Mn, Zn, Cu, and Cr) are essential for most animals because they play an important role in many physiological functions, from cell

metabolism to growth, therefore their presence and content levels are to be expected. Furthermore, their high content scarcely poses a threat to human consumers [28] and this is, probably, why there are no Maximum Levels (MLS) established for these elements.

Conversely, Cd and Pb are both non-essential elements, which can have high nephrotoxic effects [22] and they are generally considered a hazard to human health. The EU has set a limit of 0.5 ppm for the concentration in crustaceans, for both elements (EC Regulation No 1881/2006). In the present study, the content of both elements in the whole body of *P. longirostris* individuals was relatively higher or in close agreement with similar studies in the Eastern Mediterranean region (Table 3) and higher than the EU limit.

**Table 3.** Concentrations of trace elements (ppm) in shrimps from different parts of the Eastern Mediterranean Sea.

|   | Species                         | Fe          | Cu          | Zn          | Mn         | Ni        | Pb        | Cr        | Cd          | As    | Remarks                        |                            |       |       |      |                            |                           |                             |      |                            |
|---|---------------------------------|-------------|-------------|-------------|------------|-----------|-----------|-----------|-------------|-------|--------------------------------|----------------------------|-------|-------|------|----------------------------|---------------------------|-----------------------------|------|----------------------------|
| present study                             | <i>Parapenaeus longirostris</i> | 102         | 35          | 62          | 72         | 4.6       | 0.88      | 1.01      | 0.93        | 64    | Summer, wet weight, whole body |                            |       |       |      |                            |                           |                             |      |                            |
|   |                                 |             |             |             |            |           | 0.31      |           |             |       | 0.47                           | Summer, wet weight, muscle |       |       |      |                            |                           |                             |      |                            |
| Olgunoğlu et al., 2015                    | <i>Aristaeomorpha foliacea</i>  | 1.09        | 2.71        | 11.16       |            |           | 0.43      |           |             |       | Wet weight, muscle, Females    |                            |       |       |      |                            |                           |                             |      |                            |
|   |                                 |             |             |             |            |           |           |           |             |       | 24.21                          | 144.12                     | 34.79 |       | 2.5  | Wet weight, liver, Females |                           |                             |      |                            |
|   |                                 |             |             |             |            |           |           |           |             |       | 2.85                           | 3.07                       | 13.37 |       |      | 0.43                       | Wet weight, muscle, Males |                             |      |                            |
|   |                                 |             |             |             |            |           |           |           |             |       | 24.76                          | 257.88                     | 34.88 |       |      | 0.36                       | 2.71                      | Wet weight, liver, Males    |      |                            |
| Olgunoğlu, 2015                           | <i>Aristeus antennatus</i>      | 1.609       | 2.099       | 9.953       | 0.698      |           |           |           |             |       | Wet weight, Muscle             |                            |       |       |      |                            |                           |                             |      |                            |
|   | <i>Plesionika edwardsii</i>     | 0.619       | 2.040       | 5.933       |            |           |           |           |             |       | Wet weight, Muscle             |                            |       |       |      |                            |                           |                             |      |                            |
|   | <i>Plesionika martia</i>        | 0.456       | 1.357       | 4.483       |            |           |           |           |             |       | Wet weight, Muscle             |                            |       |       |      |                            |                           |                             |      |                            |
| Dökmeci et al., 2014                      | <i>Parapenaeus longirostris</i> |             | 4.16        | 12.16       |            |           | 0.22      | 1.84      | 0.13        | <0.02 | 0.86                           | Summer, wet weight, muscle |       |       |      |                            |                           |                             |      |                            |
|   |                                 |             |             |             |            |           |           |           |             |       |                                | 25.48                      | 22.42 |       | 2.23 | <0.51                      | 0.77                      | <0.05                       | 2.33 | Winter, wet weight, muscle |
|   |                                 |             |             |             |            |           |           |           |             |       |                                |                            | 3.98  | 16.17 |      | 19.25                      | 2.12                      | 0.37                        | 0.1  | 9.93                       |
| Kalogeropoulos et al., 2012               | <i>Parapenaeus longirostris</i> | 56          | 9.5         | 12          |            |           | 0.13      | 0.22      | 0.23        |       |                                | Wet Weight, muscle         |       |       |      |                            |                           |                             |      |                            |
| Pastorelli et al., 2012                   | <i>Parapenaeus longirostris</i> |             |             |             |            |           | 0.31      |           | 0.14        |       |                                | Wet weight, muscle         |       |       |      |                            |                           |                             |      |                            |
| Turkmen, 2012                             | <i>Penaeus kerathurus</i>       | 32.47       | 28.47       | 63.45       |            |           | 0.81      |           | 0.22        |       |                                | Dry weight, Muscle, Males  |       |       |      |                            |                           |                             |      |                            |
|   |                                 |             |             |             |            |           |           |           |             |       |                                | 31.26                      | 26.37 | 60.81 |      | 0.91                       | 0.18                      | Dry weight, Muscle, Females |      |                            |
| Çevik et al., 2008                        | <i>Parapenaeus longirostris</i> | 20.31       |             | 42.93       | 2.39       |           |           | 0.28      | 0.09        |       |                                | Dry weight, muscle         |       |       |      |                            |                           |                             |      |                            |
| Firat et al., 2008                        | <i>Penaeus semisulcatus</i>     | 18.69       | 32.24       | 27.75       |            |           |           | 60.38     | 16.72       |       |                                | Dry weight, muscle         |       |       |      |                            |                           |                             |      |                            |
| Gokoglou et al., 2008                     | <i>Penaeus semisulcatus</i>     | 33.89       | 6.19        | 30.84       | 0.6        |           |           |           | 2.36        |       |                                | Wet weight, muscle         |       |       |      |                            |                           |                             |      |                            |
|   | <i>Parapenaeus longirostris</i> | 11.81       | 1.33        | 14.57       | 1.52       |           |           |           | 0.23        |       |                                | Wet weight, muscle         |       |       |      |                            |                           |                             |      |                            |
| Yilmaz and Yilmaz, 2007                   | <i>Penaeus semisulcatus</i>     | 5.9 – 33.1  | 17.2 – 41.0 | 6.0 – 10.2  |            |           | 0.6 – 3.4 | 0.3 – 0.6 | 6.8 – 13.1  |       |                                | Wet weight, muscle, males  |       |       |      |                            |                           |                             |      |                            |
|   |                                 | 6.8 – 24.7  | 17.5 – 42.4 | 4.3 – 10.3  |            |           | 0.6 – 3.6 | 0.2 – 0.6 | 5.9 – 9.3   |       |                                | Wet weight, muscle, female |       |       |      |                            |                           |                             |      |                            |
| Kurun et al., 2007                        | <i>Palaemon adspersus</i>       | 32.6 – 64.7 | 30.2 – 45.3 | 25 – 70.1   | 2.9 -9.7   | 3.1 -6.4  | 2.6 – 5.9 |           | 0.10 – 0.98 |       |                                | Autumn, dry weight, muscle |       |       |      |                            |                           |                             |      |                            |
|   | <i>Palaemon serratus</i>        | 10.5 – 38.4 | 31.3 -48.7  | 21.1 – 47.1 | 0.01 – 6.8 | 2.8 – 6.7 | 2.9 – 5.2 |           | 0.70 – 1.02 |       |                                | Autumn, dry weight, muscle |       |       |      |                            |                           |                             |      |                            |
|   | <i>Parapenaeus longirostris</i> | 7.1 – 88.8  | 19.4 – 27.9 | 28.9 – 50.8 | 0.01 – 8.8 | 1.4 – 7.5 | 2.6 – 7.5 |           | 0.53 – 1.22 |       |                                | Autumn, dry weight, muscle |       |       |      |                            |                           |                             |      |                            |
| Satsmadjis and Voutsinou-Taliadouri, 1983 | <i>Parapenaeus longirostris</i> | 7.6         | 6.5         | 13.3        | 0.62       | 0.06      | 0.28      | 0.05      | 0.05        |       |                                | Autumn, Wet Weight, muscle |       |       |      |                            |                           |                             |      |                            |
| Balkas et al., 1982                       | <i>Penaeus kerathurus</i>       | 3.1         | 7.4         | 13.2        | 0.2        | 1.4       | 0.34      | 0.14      | 0.03        |       |                                | Wet weight, muscle         |       |       |      |                            |                           |                             |      |                            |

These elevated levels could be attributed to the fact that the surface sediments of Pagasitikos Gulf have displayed elevated levels of Pb [29-31] and also, because whole-body samples include the hepatopancreas and viscera, where cadmium and lead burdens are, usually, greater [22, 23].

A key aspect in the assessment of the risks to human health from potentially harmful chemicals in food is the knowledge of the dietary intake of such substances, which must remain within determined safety margins [25]. In the present study, according to the EWI values, the dietary intake of Cd and Pb from *P. longirostris* seems to be higher when the whole body is consumed compared to the muscle tissue results. Still, in both cases, that intake represents the 11.9 – 23.6% and 0.8 – 2.2% of

the PTWI for Cd and Pb, respectively, which is an indication that the health risk posed by the consumption of local *P. longirostris* shrimps can be considered insignificant. In addition to the interpretation of the weekly intake results, the estimated THQs were very low for both elements (Cd and Pb) and in both cases (whole body and muscle tissue). According to the guideline for interpreting hazard quotients (HQ) calculations, for HQ values lower than 0.1 no hazard exists, while for values between 0.1–1.0 the hazard is low [32]. Therefore, from the estimated TTHQ values of 0.058 and 0.050 for the whole body and muscle tissue, respectively, it could be assumed that the health risk from the consumption of *P. longirostris* shrimps from the Pagasitikos Gulf is minimal.

**Table 4.** Estimated weekly intakes (EWI, mg kg<sup>-1</sup> body weight), percent coverage of provisional tolerable weekly intakes (PTWI), target hazard quotients (THQ), and total hazard quotient (TTHQ) for the trace elements Cd and Pb in the whole body and muscle tissue of *P. longirostris*.

| Element  | Whole Body |       | Muscle Tissue |       |
|----------|------------|-------|---------------|-------|
|          | Cd         | Pb    | Cd            | Pb    |
| EWI      | 0.59       | 0.56  | 0.30          | 0.20  |
| PTWI (%) | 23.6       | 2.2   | 11.9          | 0.8   |
| THQ      | 0.046      | 0.013 | 0.043         | 0.007 |
| TTHQ     | 0.058      |       | 0.050         |       |

To conclude, it is worth noting that the findings of the present study were, in general, comparable to those obtained by other studies in the Eastern Mediterranean region, either in the same species or in other shrimps. In some cases, the high content values recorded, could be interpreted as signs of anthropogenic influence, but still, none of the elements, including the hazardous Cd and Pb, seem to pose a threat to human health. Thus, the deep-water rose shrimp from Pagasitikos Gulf could be considered safe for consumption, probably with a piece of general precautionary advice to the public to avoid consuming anything else than the muscle tissue, because the exposure to trace elements is the lowest, when compared with the whole body.

#### ACKNOWLEDGMENTS

The authors would like to thank the skipper and crew of the vessel “Panagiotis – Anestis” for their valuable contribution to the bottom-trawl expedition. They would also like to extend their gratitude to the anonymous reviewers for their constructive comments and suggestions.

#### Conflict of interests

We know of no conflicts of interest associated with this publication, and there has been no significant financial support for this work that could have influenced its outcome.

#### REFERENCES

1. Wang H., Liang Y., Li S., Chang J., 2013. Acute Toxicity, Respiratory Reaction, and Sensitivity of Three Cyprinid Fish Species Caused by Exposure to Four Heavy Metals. PLOS ONE. 8(6), e65282. doi: 10.1371/journal.pone.0065282.
2. Saha N., Mollah M.Z.I., Alam M.F., Safiur Rahman M., 2016. Seasonal investigation of heavy metals in marine fishes captured from the Bay of Bengal and the implications for human health risk assessment. Food Control. 70110-118. doi: 10.1016/j.foodcont.2016.05.040.
3. Ghaneian M.T., Bhatnagar A., Ehrampoush M.H., Amrollahi M., Jamshidi B., Dehviri M., Taghavi M., 2017. Biosorption of hexavalent chromium from aqueous

- solution onto pomegranate seeds: kinetic modeling studies. *Int J Environ Sci Te.* 14(2), 331-340. doi: 10.1007/s13762-016-1216-8.
4. Bosch A.C., O'Neill B., Sigge G.O., Kerwath S.E., Hoffman L.C., 2016. Heavy metals in marine fish meat and consumer health: a review. *J Sci Food Agric.* 96(1), 32-48. doi: 10.1002/jsfa.7360.
5. Abdennour C., Smith B.D., Boulakoud M.S., Samraoui B., Rainbow P.S., 2000. Trace metals in marine, brackish and freshwater prawns (Crustacea, Decapoda) from northeast Algeria. *Hydrobiologia.* 432(1), 217-227. doi: 10.1023/A:1004027204088.
6. Sonone S.S., Jadhav S., Sankhla M.S., Kumar R., 2021. Water contamination by heavy metals and their toxic effect on aquaculture and human health through food Chain. *Lett. Appl. NanoBioScience.* 10(2), 2148-2166. doi: 10.33263/LIANBS102.21482166.
7. Leonart J., Maynou F., 2003. Fish stock assessments in the Mediterranean: state of the art. *Sci Mar.* 67(S1), 37-49. doi: 10.3989/scimar.2003.67s137.
8. Colloca F., Mastrantonio G., Lasinio G.J., Ligas A., Sartor P., 2014. *Parapenaeus longirostris* (Lucas, 1846) an early warning indicator species of global warming in the central Mediterranean Sea. *J Marine Syst.* 13829-39. doi: 10.1016/j.jmarsys.2013.10.007.
9. Phillips D.J.H., Rainbow P.S., 1993. The Biomonitoring of Trace Metals and Radionuclides. In: *Biomonitoring of Trace Aquatic Contaminants*, Springer Netherlands: Dordrecht. pp. 79-132.
10. Kalogeropoulos N., Karavoltos S., Sakellari A., Avramidou S., Dassenakis M., Scoullou M., 2012. Heavy metals in raw, fried and grilled Mediterranean finfish and shellfish. *Food Chem Toxicol.* 50(10), 3702-3708. doi: 10.1016/j.fct.2012.07.012.
11. Sobrino I., Silva C., Sbrana M., Kapiris K., 2005. A review of the biology and fisheries of the deep water rose shrimp, *Parapenaeus longirostris*, in European Atlantic and Mediterranean waters (Decapoda, Dendrobranchiata, Penaeidae). *Crustaceana.* 78(10), 1153-1184. doi: 10.1163/156854005775903564.
12. SoHelFi, 2007. *State of Hellenic Fisheries*, C. Papaconstantinou, Editors. Hellenic Centre for Marine Research. p. 466.
13. Kapiris K., Markovic O., Klaoudatos D., Djurovic M., 2013. Contribution to the biology of *Parapenaeus longirostris* (Lucas, 1846) in the South Ionian and South Adriatic Sea. *Turk J Fish Aquat Sc.* 13(4), 647-656. doi: 10.4194/1303-2712-v13\_4\_10.
14. Fortibuoni T., Bahri T., Camilleri M., Garofalo G., Gristina M., Fiorentino F., 2010. Nursery and Spawning Areas of Deep-water Rose Shrimp, *Parapenaeus longirostris* (Decapoda: Penaeidae), in the Strait of Sicily (Central Mediterranean Sea). *J Crustacean Biol.* 30(2), 167-174. doi: 10.1651/09-3167.1.
15. Fellah H., Maouel D., Hemida F., 2021. Research Article: Population dynamics of *Parapenaeus longirostris* (Decapoda: Penaeidae) (Lucas, 1846), in Central Algerian waters (South-Western of Mediterranean Sea). *Iran J Fish Sci.* 20(5), 1426-1441. doi: 10.22092/ijfs.2021.125054.
16. Pastorelli A.A., Baldini M., Stacchini P., Baldini G., Morelli S., Sagratella E., Zaza S., Ciardullo S., 2012. Human exposure to lead, cadmium and mercury through fish and seafood product consumption in Italy: a pilot evaluation. *Food Addit. Contam.* 29(12), 1913-1921. doi: 10.1080/19440049.2012.719644.
17. Satsmadjis J., Voutsinou-Taliadouri F., 1983. *Mytilus galloprovincialis* and *Parapenaeus longirostris* as bioindicators of heavy metal and organochlorine pollution. *Mar Biol.* 76(2), 115-124. doi: 10.1007/BF00392728.
18. Dökmeçi A.H., Yildiz T., Ongen A., Sivri N., 2014. Heavy metal concentration in deepwater rose shrimp species (*Parapenaeus longirostris*, Lucas, 1846) collected from the Marmara Sea Coast in Tekirdağ. *Environ Monit Assess.* 186(4), 2449-2454. doi: 10.1007/s10661-013-3551-2.
19. Çevik F., Bayhan Y.K., Derici O.B., 2008. Metal Concentrations in the Muscle of Male and Female Shrimp (*Parapenaeus longirostris* Lucas, 1846) Collected from Marmara Sea and Their Relationships with Season. *Asian J Chem.* 20(3), 2229-2237.
20. Yılmaz A.B., Yılmaz L., 2007. Influences of sex and seasons on levels of heavy metals in tissues of green tiger shrimp (*Penaeus semisulcatus* de Hann, 1844). *Food Chem.* 101(4), 1664-1669. doi: 10.1016/j.foodchem.2006.04.025.
21. Traina A., Bono G., Bonsignore M., Falco F., Giuga M., Quinci E.M., Vitale S., Sprovieri M., 2019. Heavy metals concentrations in some commercially key species

- from Sicilian coasts (Mediterranean Sea): Potential human health risk estimation. *Ecotox Environ Safe.* 168466-478. doi: 10.1016/j.ecoenv.2018.10.056.
22. Eisler R. 2010. *Compendium of Trace Metals and Marine Biota*, Elsevier: Oxford, UK.
23. Pourang N., Dennis J.H., Ghourchian H., 2004. Tissue Distribution and Redistribution of Trace Elements in Shrimp Species with the Emphasis on the Roles of Metallothionein. *Ecotoxicology.* 13(6), 519-533. doi: 10.1023/B:ECTX.0000037189.80775.9c.
24. Vlachonikolis I.G., Marriott F.H.C., 1995. Evaluation of censored contamination data. *Food Addit Contam.* 12(5), 637-644. doi: 10.1080/02652039509374352.
25. Storelli M.M., 2008. Potential human health risks from metals (Hg, Cd, and Pb) and polychlorinated biphenyls (PCBs) via seafood consumption: Estimation of target hazard quotients (THQs) and toxic equivalents (TEQs). *Food Chem Toxicol.* 46(8), 2782-2788. doi: 10.1016/j.fct.2008.05.011.
26. Zaza S., de Balogh K., Palmery M., Pastorelli A.A., Stacchini P., 2015. Human exposure in Italy to lead, cadmium and mercury through fish and seafood product consumption from Eastern Central Atlantic Fishing Area. *J Food Compos Anal.* 40148-153. doi: 10.1016/j.jfca.2015.01.007.
27. Bastami K.D., Afkhami M., Mohammadizadeh M., Ehsanpour M., Chambari S., Aghaei S., Esmailzadeh M., Neyestani M.R., Lagzaee F., Baniamam M., 2015. Bioaccumulation and ecological risk assessment of heavy metals in the sediments and mullet *Liza klunzingeri* in the northern part of the Persian Gulf. *Mar Pollut Bull.* 94(1-2), 329-334. doi: 10.1016/j.marpolbul.2015.01.019.
28. Fraga C.G., 2005. Relevance, essentiality and toxicity of trace elements in human health. *Mol. Aspects Med.* 26(4-5), 235-244. doi: 10.1016/j.mam.2005.07.013.
29. Voutsinou-Taliadouri F., Satsmadjis J., 1982. Trace metals in the Pagassitikos Gulf, Greece. *Estuar. Coast. Shelf S.* 15(2), 221-228. doi: 10.1016/0272-7714(82)90029-4.
30. Voutsinou-Taliadouri H.F., Georgakopoulou-Grigoriadou E., 1989. Heavy metal concentrations in surface sediments from Pagassitikos Gulf, Greece. *Toxicol Environ Chem.* 20-21(1), 53-58. doi: 10.1080/02772248909357359.
31. Karageorgis A.P., Sioulas A.I., Anagnostou C.L., 2002. Use of surface sediments in Pagassitikos Gulf, Greece, to detect anthropogenic influence. *Geo-Mar Lett.* 21(4), 200-211. doi: 10.1007/s00367-001-0086-2.
32. Lemly A.D., 1996. Evaluation of the Hazard Quotient Method for Risk Assessment of Selenium. *Ecotox Environ Safe.* 35(2), 156-162. doi: 10.1006/eesa.1996.0095.

