# Preconcentration of Hg(II) with novel nano-Fe<sub>3</sub>O<sub>4</sub>- sorbents prior to determination by CVAAS

# A. Moghimi

Department of Chemistry, Varamin (Pishva) Branch, Islamic Azad University, Varamin, Iran

Received: 26 Jun 2016; Accepted: 29 August 2016

**ABSTRACT:** A novel, simple, sensitive and effective method has been developed for preconcentration of Hg(II) on nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED solid-phase extraction adsorbent. In this paper, Dioctyl phthalate (DOP) was used to encapsulate nano-Fe<sub>3</sub>O<sub>4</sub> and produce a nano-Fe<sub>3</sub>O<sub>4</sub>-DOP sorbent based new sorbent was prepared. This was treated with ethylenediamine (ED) in another solvent-free procedure for the formation of a novel nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED. Flame atomic absorption spectrometer was utilized for determination of Hg(II). Some of the important parameters on the preconcentration and complex formation were selected and optimized. Under the optimized conditions the limit of detection (LOD) and limit of quantification (LOQ) were 49.6, 163.7 ng L<sup>-1</sup> and the proposed method has a good reproducibility 0.90% (RSD %) and preconcentration factor was found to be 200. The methodology was applied for determination of Hg(II) in natural water samples and satisfactory results were obtained.

**Keywords:** Cadmium; Central Composite Design; Dioctyl phthalate (DOP) was used to encapsulate nano-Fe<sub>3</sub>O<sub>4</sub>; Flame Atomic Absorption Spectrometer; Preconcentration; Water

# INTRODUCTION

The properties of nano magnetic particles research on their functionality as magnetic carrier, (Bulut, *et al.*, 2010, Graf, *et al.*, 2007), also in medical imaging uses such as magnetic resonance imaging (MRI) (Ansari, *et al.*, 1999, Margel, *et al.*, 1997), in isolating substances and materials in anions– cations pre-concentration (Afkhami and Moosavi, 2010, White, *et al.*, 2009, Tuutijarvi, *et al.*, 2009, Batterham, *et al.*, 1997, Hummers and Offeman, 1958, Camel, 2003). The determination of Hg is usually carried out by flame and graphite furnace atomic absorption spectrometry (AAS) (Castillo, *et al.*, 2005, Eaton, *et al.*, 1995) as well as spectrometric methods (Welcher and Boschmann, 1979, Marczenko, 1986). However, due to the presence of Hg in

(\*) Corresponding Author - e-mail: alimoghimi@iauvaramin.ac.ir

medicinal and environmental samples at low levels, its separation from other elements presents and also the use of a preconcentration step prior to its determination is usually necessary. Different methods, co-extractant ligands have attracted considerable attention (Carasek, *et al.*, 2002, Ceccarini, *et al.*, 2005). However, the use of classical extraction methods for this purpose is usually for the Extraction and separation of Hg ions have been suggested including liquid chromatography (Tuzen, 2009) supercritical fluid extraction (Choi and Choi, 2003, Tuzen, *et al.*, 2004, Karousis, *et al.*, 2011, Smith and March, 2001, Mermoux, *et al.*, 1991, Cataldo, 2003), flotation (Tohidifar, *et al.*, 2013), aggregate film formation (Thistlethwaite and Hook, 2000), liquid membrane (Taguchi, *et al.*, 1997), column adsorption of pyrocatechol violet-Hg complexes on activated carbon (Tajodini and Moghimi, 2010), ion pairing (Thistlethwaite and Hook, 2000), ion pairing (Stevens, et al., 2007), preconcentration with yeast (Su, 2003), and solid phase extraction using C<sub>18</sub> cartridges and disks (Su, et al., 2003, Soylak, et al., 2003, Szabo, et al., 2005, Moghimi, 2014). Solid phase extraction (SPE) or liquid-solid extraction is poplar and growing techniques that are used to sample preparation for analysis. However, the disks modified ligand is a selective approach to separation and pre- concentration of heavy metals in water samples (Moghimi A., Abdouss, 2013, Moghimi, et al., 2007, Tarigh and Shemirani, 2013, Moghimi and Poursharifi, 2012, Moghimi, 2014, Moghimi and Siahkalrodi, 2013, Moghimi and Yari, 2014, Moghimi and Akbarieh, 2014, Moghimi, 2013). In a recent series of papers, (Moghimi, et al., 2012, Moghimi, et al., 2013, Moghimi and Shabanzadeh, 2012) the determinate relies on the esterase activity of a DNA-linked Hg complex. For optimization of the system and exploration of structure- activity relationships, a sensitive probe would be useful, which allows straightforward detection of esterase activity of ligated Hg<sup>2+</sup> in low concentration. The chelated ions were desorbed and determined by CVAAS. The modified solid phase could be used at least 50 times with acceptable reproducibility without any change in the composition of the sorbent, nano-Fe<sub>2</sub>O<sub>4</sub>-DOP-ED. In the present work, nano-Fe<sub>2</sub>O<sub>4</sub>-DOP-ED was employed for production of solid phase. The synthesized and characterized new sorbent nano-Fe<sub>2</sub>O<sub>4</sub>-DOP-ED was utilized for preconcentration of Hg(II) from water samples. Determination of Hg(II) concentration was achieved by CVAAS after preconcentration procedure.

# MATERIALS AND METHODS

#### **Reagents and Chemicals**

The analytical grade deionized water was obtained by reverse osmosis system. All containers and glassware were kept overnight in 10% nitric acid and rinsed three times with water before use.1,000 mg L<sup>-1</sup> stock Hg standard solution was prepared from Hg(NO<sub>3</sub>)<sub>2</sub>.4H<sub>2</sub>O (Merck) and diluted as required to the  $\mu$ g L<sup>-1</sup> levels.

BDH Limited, Poole, England. Dioctyl phthalate, DOP, (purity> 99.55%), sodium hydroxide (NaOH) and ethylenediamine (ED) was purchased from BDH, UK. Hg nitrate, Hg(NO<sub>3</sub>)<sub>2</sub>, cadmium nitrate, Cd(NO<sub>3</sub>)<sub>2</sub> and NiCl<sub>2</sub>-6H<sub>2</sub>O was purchased from Merck. In interference study, cations were added as nitrates and the anions were added as sodium salts. Feasibility of the suggested method was tested with Lake Ontario water certified reference material (TMDA-53.3, lot 0310).

#### Synthesis of magnetic nano- $Fe_3O_4$ sorbent

The magnetic nano-iron oxide was first synthesized according to a previously reported method (Graf, *et al.*, 2007). A 6.1 g sample of FeCl<sub>3</sub>-6H<sub>2</sub>O and 4.2 g FeSO<sub>4</sub>-7H<sub>2</sub>O were dissolved in 100 mL of distilled water. A total of 25.0 mL of 6.5 M-NaOH was slowly added to the above solution. The reaction mixture was stirred for 4 h using a magnetic stirrer. After the complete addition of NaOH, the formed black precipitate, nano-FejOa, was then washed several times with distilled water, collected in a pure form by the assistance of an external magnetic field and dried in an oven at 70°C.

# Synthesis of nano-Fe<sub>3</sub>O<sub>4</sub>-DOP sorbent

The Fe<sub>3</sub>O<sub>4</sub>-encapsulated-DOP nano-sorbent was prepared by the addition of 20.0 mL of DOP to 10.0 g of nano-Fe<sub>3</sub>O<sub>4</sub> sorbent. This mixture was heated under stirring at 80-90 °C for 3 hands the product nano-Fe<sub>3</sub>O<sub>4</sub>-DOP sorbent was filtered, washed with methanol, collected in a pure form by a magnetic field and dried in an oven at 70 °C until complete dryness.

# Synthesis of magnetic nano- $Fe_3O_4$ -DOP- ED sorbent

A 10.0 g sample of nano-Fe<sub>3</sub>O<sub>4</sub>-DOP addition to 5 mL of ED and these two reactants were combined together by heavy grinding and mixing in a mortar for 5 h. The produced nano- Fe<sub>3</sub>O<sub>4</sub> DOP-TETA sorbent was heated to dryness in an oven at 60 °C.

## Apparatus

A Philips X Pert-Pro diffracrometer (Hg K $\alpha\lambda$ = 1.54060 A°, 30 mA, 40 kV), and Perkin Elmer Spectrum 65 FTIR-ATR spectrometer were used to confirm the synthesized Fe<sub>3</sub>O<sub>4</sub>-encapsulated-DOP nano-sorbent.

Determination of Hg(II) in solutions were carried out by Perkin Elmer AAnalyst200 CVAAS equipped with deuterium background correction. All measurements were performed in an air/acetylene flame. GFL 3005 orbital shaker having speed and time control was used for preparation of the sorbent. During the solid phase extraction experiments, VelpScientifica SP311 peristaltic pump with Tygon tubes was used. A Thermo Orion 5 Star model pH meter, Heidolph MR 3001 K model magnetic stirrer, Sartorius TE214S electronic balance, Eppendorf Research micro pipettes were used for the present work. Funnel tipped glass tube (10x100 mm) equipped with stopcock was used as a column for the preconcentration experiments.

# **Preparation of Solid Phase**

Sorption characteristics of Hg(II) ions by magnetic nano-sorbents by the batch equilibrium technique. The applicability of magnetic nano-  $Fe_3O_4$ -DOP-ED sorbent for extraction of Hg(II) ions was studied by the batch equilibrium technique under several experimental controlling factors. These include the effect of pH, contact time, sorbent dosage, initial metal ion concentration and interfering ions.

#### General procedure

100 mL solution with 50 ngmL<sup>-1</sup> Hg content was prepared as a sample. This solution was added to the magnetic nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED sorbent and by using buffer solution, its pH was modified to 3.0, followed the container was shaken for 25 min in order to allow easier Hg ions absorption of the nano- Fe<sub>3</sub>O<sub>4</sub>-DOP-ED sorbent (Tarigh and Shemirani, 2013). At the end of this stage, the nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED sorbent which had been formed on the bottom layer of the beaker was removed through applying external field with1.4T magnetic powers and was immediately decanted outside the supernatant. 1 mL of 1 molL<sup>-1</sup> ion of eluent was injected into CVAAS for measurement of Hg(II) ions concentrations.

# **RESULTS AND DISCUSSION**

# Surface morphology

Figs. 1a and b show the morphology and size of mag-







Fig. 1. Images of (a) Scanning Electron Microscopy (SEM) and (b) Transmission Electron Microscopy (TEM) of magnetic nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED

netic nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED sorbent as shown in Figs. 1a and b. The particles of nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED 339 sorbent retained a homogeneous distribution in the range of 340 5.0–20.0 nm.

# Effect of pH

The pH of the sample solution plays an important role in retention of metals on sorbent. The pH of the model solutions containing  $5\mu g$  Hg(II) were adjusted to certain value using diluted HNO<sub>3</sub> and NaOH. As shown



Fig. 2. Influence of sample pH and dissolving solvent of nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED on the percentage recovery of Hg(II).

	Fo	For sorption Levels			evels	-	ŀ	For elu	ition	Leve	ls
Factors	-α	-1	0	+1	$+\alpha$	Factors	-0	ι -1	0	+1 -	+α
pH	3.3	4	5	6	6.7	Flow rate (mL min <sup>-1</sup> )	2.3	3	4	5	5.7
						Eluent Concentration					
Flow rate (mL min <sup>-1</sup> )	2.3	3	4	5	5.7		0.08	0.25	0.5	0.75	0.92
						$(mol L^{-1})$					
Sample volume (mL)	8.0	25	50	75	92.0	Eluent Volume	3.3	4	5	6	6.7

Table 1. Levels and the real values of factors utilized in CCD.

in Fig. 2, the recovery results were not dramatically affected by the change in pH between 4 and 7. According to this, pH= 3.0 were chosen as centre value for the optimization procedure.

# **Desorption reagent**

0.5 mol L<sup>-1</sup> of HNO<sub>3</sub>, HCl, CH<sub>3</sub>COOH, H<sub>2</sub>SO<sub>4</sub> and H<sub>2</sub>O<sub>2</sub> were tested for desorption of Hg(II) from nano-Fe<sub>3</sub>O<sub>4</sub> -DOP- ED. The recovery percentages were varied between 5.7-87.9 % except HNO<sub>3</sub> elution ex-

Table 2.	Experiments	and	recovery	values	for	the	sorption
of Hg(II)							

	v	X <sub>2</sub>	$\mathbf{X}_{3}$	Decovery
Run	<b>л</b> <sub>1</sub>	Flow rate	Sample volume	(Q())
	рН	(mL min <sup>-1</sup> )	(mL)	(%)
1	4	3	25	99.1
2	6	3	25	102
3	4	5	25	98.6
4	6	5	25	101.4
5	4	3	75	99.6
6	6	5	75	99.4
7	4	5	75	105
8	6	5	75	98.6
9	5	4	50	102.4
10	3.3	4	50	96.6
11	6.7	4	50	97.1
12	5	2.3	50	102.7
13	5	5.7	50	98.4
14	5	4	8.0	101.6
15	5	4	92.0	99.0
16	5	4	50	99.8
17	5	4	50	101.5
18	5	4	50	98.8
19	5	4	50	101.0
20	5	4	50	99.7

Table 3. Experiments and recovery values for the elution of Hg(II)

	v	X <sub>2</sub>	X <sub>3</sub>	Recovery
Run	л <sub>1</sub>	Flow rate	Sample volume	
	рн	(mL min <sup>-1</sup> )	(mL)	(%)
1	3	0.25	4	99.6
2	5	0.25	4	103
3	3	0.75	4	98.5
4	5	0.75	4	101.6
5	3	0.25	6	99.7
6	5	0.25	6	99.8
7	3	0.75	6	101.6
8	5	0.75	6	99.6
9	4	0.5	5	99.7
10	2.3	0.5	5	101.8
11	5.7	0.5	5	97.1
12	4	0.08	5	1007
13	4	0.92	5	98.8
14	4	0.5	3.3	101.7
15	4	0.5	5	99.8
16	4	0.5	5	99.1
17	4	0.5	5	101.1
18	4	0.5	5	98.4
19	4	0.5	5	100.5
20	4	0.5	5	99.0

periments. It was observed that  $HNO_3$  is the best as an eluent and the recovery percentage is  $95.5\pm1.5$  %.

# Effect of flow rate on sorption and elution

The retention of a metal ion on the sorbent also depends on the flow rate of the sample solution. Thus, the both effect of flow rate of the sample and eluent on the sorption and desorption of Hg ions were investigated between 3-20 mL min<sup>-1</sup>. Quantitative results (> 95%) were obtained up to 10 and 8 mL min<sup>-1</sup> for

#### Int. J. Bio-Inorg. Hybr. Nanomater., 5(3): 203-212, Autumn 2016

<b>Optical condition</b>			
sorption		elution	
pН	2.9	Flow rate (mL min <sup>-1</sup> )	4.1
Flow rate (mL min <sup>-1</sup> )	4.3	Eluent concentration (mol L <sup>-1</sup> )	0.6
Sample volume (mL)	55	Eluent volume (mL)	5.1

Table 4. Optimum preconcentration conditions for Hg(II) by nano- Fe3O4 -DOP- ED

sorption and elution, respectively. In order to avoid an abrupt change in adsorption and increase the contact time of the sample solution with the sorbent, flow rate was selected as 4 mL min<sup>-1</sup> for sorption and elution.

# **Optimization of variables**

Optimization of the preliminary studies was achieved by a three level full factorial CCD with 20 runs. Table 1 lists the maximum, minimum and centre values of the variables for sorption and elution. The metal concentration in solutions was determined with external standard calibration method by CVAAS. The experimental design matrix and the results for sorption and elution are given in Tables 2 and 3. Quadratic Eq. (1) and Eq. (2), which were obtained by using data of response values, are given below for sorption and elution, respectively.

$Y = 0.512150 - 0.72565 x_1 - 0.70169 x_2 - 0.98326 x_3 + +0.05653$	$83x_1^2$
+0.171215 $x_2^2$ + 0.493058 $x_3^2$ + 1.565734 $x_1x_2$ + 1.5556039 $x_1x_3$	
$+1.606564 x_{2}x_{3}$	(1)

$$\begin{split} Y &= \ 0.683345 - 0.018065 \ x_1 - 1.220286 \ x_2 - 0.130547 \ x_3 + 0.576546 \ x_1^2 \\ &+ 1.235334 \ x_2^2 + 0.33238 \ x_3^2 + 0.071217 \ x_1 x_2 + 0.06884 \ x_1 x_3 \\ &+ 0.07096 \ x_2 x_3 \end{split}$$

 $x_1$ ,  $x_2$  and  $x_3$  represents the 1st, 2nd and 3rd factors. Derivatives of the equations in terms of  $x_1$ ,  $x_2$  and  $x_3$ were equalized to zero and solved using Microsoft<sup>®</sup> Excel. The real values presented in Table 4 obtained from CCD and used as optimal conditions for further experiments.

# Interference effects

The preconcentration procedures of trace metal ions can be strongly affected by other ions. For this reason,

Diverse ion	Amounts taken (mg)	% Found	%Recovery of Hg <sup>2+</sup> ion
Na <sup>+</sup>	92.4	1.15(2.4) <sup>b</sup>	98.7(1.8)
$K^+$	92.5	1.32(2.3)	98.5(2.9)
$Mg^{2+}$	14.5	0.7(1.2)	98.9(1.8)
$Ca^{2+}$	26.3	2.25(3.0)	98.5(1.6)
$\mathrm{Sr}^{2+}$	2.45	2.85(2.5)	98.4(2.0)
$Ba^{2+}$	2.66	3.16(2.1)	98.3(2.3)
$Mn^{2+}$	2.66	1.75(2.2)	97.3(2.8)
$\mathrm{Co}^{2^+}$	2.16	1.4(2.3)	99.1(2.9)
Ni <sup>2+</sup>	1.65	2.0(2.4)	98.5(2.6)
$Zn^{2+}$	2.78	1.97(2.1)	98.4(2.2)
$Cd^{2+}$	2.55	1.92.0)	98.2(2.8)
$Pb^{2+}$	0.54	2.7(1.9)	97(2.7)
$\mathrm{Ag}^{\scriptscriptstyle +}$	2.63	3.45(2.9)	96.6(2.9)
$Cr^{3+}$	1.73	2.92(2.30	97.3(2.4)
$\mathrm{UO}^{2^+}$	2.84	2.8(2.1)	98.3(2.7)

Table 5.	Separation	of Ha	from	binarv	mixtures <sup>*</sup>
10010 0.	ooparation	01119		Sincery	1111/100

<sup>a</sup> Initial samples contained 10µg Pb<sup>2+</sup> and different amounts of various ions in 100 mL water (0.1 M acetate ion).

<sup>b</sup> Values in parentheses are RSDs based on five individual replicate analysis.

#### A. Moghimi

Table 6. Analysis of standard reference material

рН	Certified value	Found value	Recovery%
CRM (TMDA-53.3)	118.0±5.5	125±8.6	98.8

Table 7. Recovery of Pb added to 1000mL of different water samples (containing 0.1Macetate at pH= 3.0).

Sample	Pb <sup>2+</sup> added (µg)	<b>Pb<sup>2+</sup> determined</b> (ng.mL <sup>-1</sup> )	ICP-AES
T	0.0	$1.74(1.4)^{a}$	ND
Tap water	10.0	11.98(3.2)	11.7
C	0.0	4.45(2.4)	ND
Snow water	10.0	14.97(2.0)	14.7
Rain water	0.0	2.65(2.3)	ND
	10.0	12.75(2.4)	12.3
Sea Water	0.0	12.64(2.3)	12.5
	10.0	22.93(2.0)	23.1

<sup>a</sup> Values in parentheses are %RSDs based on five individual replicate analysis

<sup>b</sup> Not detected

the effects of matrix ions were investigated under optimal conditions. The results revealed that the nano- $Fe_3O_4$ -DOP-ED behaves as a neutral ionosphere in the pH range 3.0 (Moghimi, 2012, Moghimi, 2007) so that the Hg ions are retained as ion pair complexes by the membrane disks. As seen, acetate ion is the most efficient counter anion for the SPE of Hg(II) ions. The influence of the concentration of sodium acetate ion on Hg recovery was investigated, and the results are shown in Table 4.

As seen, the percent recovery of Hg<sup>2+</sup> increased with the acetate concentration until a reagent concentration of about 0.1 M is reached, beyond which the recovery remained quantitative. Moreover, acetate ion acts as a suitable buffering agent, while it effectively contributes to the ions- pair formation; thus, in the SPE experiments, there was no need for the addition of any buffer solution.The tolerance limit of coexisting ions was given in Table 5.The experiments indicated that, no further sample treatment or masking reagents are needed.

# Effect of sample volume

The preconcentration studies were applied to solutions within the range of 25-1000 mL containing 5  $\mu$ g amount of Hg(II) to explore the possibility of enriching at low concentration with high enrichment factor. The recovery value was obtained as 98.7 % at 1000 mL sample volume by analysing 5 mL eluate and the highest preconcentration factor was found to be 200.

# Analytical figures of merit

The accuracy and precision (RSD %) of the proposed solid phase extraction procedure underoptimal conditions were investigated (n= 10) as  $102.0\pm0.2$  % and 2, respectively. Limits of detection (LOD) and the limits of quantification (LOQ) were obtained by using a criterion signal-to-noise ratio of 3 and 10, respectively. The results were calculated 49.6 ng L<sup>-1</sup> for LOD and 163.7 ng L<sup>-1</sup> for LOQ.

# Validation and application of the improved method

The developed procedure was validated by Hg (II) determination in certified reference material. The results are given in Table 6. A statistical evaluation was performed by Student's t test and t value was calculated as 2.95. Critical t value (4.30) is higher than the calculated one at 95 % confidence level. This test showed no significant difference between Hg(II) concentration obtained from the presented method and the certified value.

Fluert	Preconcentration	Mathad	Chelating agent/solid phase	Defense
Liuent	factor	Method	adsorbent	Kelerences
				(Moghimi and
$7 \text{ mol } L^{-1} \text{ HCl}$	200	CVAAS	Dithizone/microcrystalline naphthalene	Shabanzadeh,,
				2012)
$0.1 \text{ mol } L^{-1} \text{ HNO}_3$	50	Atomic absorption	$\beta$ -Naphthol/polyurethane foam	(Xie, et al., 2008)
		spec- trophotometry		
$10 \text{ mol } L^{-1} \text{HCl}$	200	CVAAS	Dithizone/silica gel	(Goswami, <i>et al.,</i> 2003)
Tetraphenyl- borate	80	Anodic stripping voltammetry	$HgI_4^{2-}$ -Aliquat-336/naphthalene	(Moghimi, 2012)
$H_2SO_4 - H_2O_2$ mixture	40	ICP-AES	DuoliteGT-73 resin	(Moghimi, 2007)
Water	5	CVAAS	Dithioacetal/SiO <sub>2</sub>	(McAllister, <i>et al.</i> , 2007)
1 mol L <sup>-1</sup> HBr	50	CVAAS	Hexathia18 crown-6 tetraone/Empore disk	(Pourreza, <i>et al.,</i> 2012)
1 mol L <sup>-1</sup> HBr	100	CVAAS	1,5-Diphenylcarbazone/SDS coated alumina	(Ekinci Dogan and Akcin, 2007)
HNO <sub>3</sub>	_	ICP-AES	1,5-Bis(2-pyridyl)-3-sulfophenylmethylene) thiocarbonohydrazide/Dowex anion	(Moghimi, 2014)
			exchange resin	
6 mol L <sup>-1</sup> HCl	50	CVAAS	4-(2-Pyridyl azo) resorcinol/nano-sized SiO <sub>2</sub>	(Ghaedi, <i>et al.,</i> 2006)
2 mol L <sup>-1</sup> HNO <sub>3</sub> in acetone	300	CVAAS	2-Mercaptobenzoxazole/chromosorb	(Ensafi, <i>et al.</i> , 2003)
HNO <sub>3</sub>	200	CVAAS	nano- Fe <sub>3</sub> O <sub>4</sub> -DOP- ED	Present work

Table 8. Comparison with other solid phase adsorbents

# Real sample analysis

To assess the applicability of the method to real samples, it was applied to the extraction and determination of Hg from different water samples. Tap water (Tehran, taken after 10 min operation of the tap), rain water (Tehran, 20 January, 2015), Snow water (Varamin, 6 February, 2015) and Sea water (taken from Caspian Sea, near the Mahmoud-Abad shore) samples were analyzed (Table 7). As can be seen from Table 4 the added Hg ions can be quantitatively recovered from the water samples used. As is seen, the recovered Hg ion reveals that the results are quite reliable and are in satisfactory agreement with those obtained by IC-PAES (Table 7).

# CONCLUSIONS

In the present study, a new sorbent is prepared by using nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED and for preconcentration of Hg(II) from natural water samples. With nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED was prepared easily and sorbed Hg(II) rapidly. The preconcentration parameters pH, flow rate and sample volume for sorption procedure was obtained by CCD as 5.2, 4.3 mL min<sup>-1</sup>, 55.0 mL, respectively. Elution parameters, flow rate, eluent concentration and eluent volume were also obtained as 4.1 mL min<sup>-1</sup>, 0.6 mol L<sup>-1</sup> and 5.1 mL, respectively. Enrichment factor was found to be 200 when 1000 mL of water sample (included 5 µg) passed through the nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED column. The interference effects of some ions were also investigated according to the improved method. The tolerance limits were between750-10,000 times greater as can be seen in Table 5. The mean recovery values for spiked water samples were satisfactory and confirmed the validity of the method. Additionally, mean %RSD value was 2 and showed that the precision of the method is quite good. Preconcentration factor, LOD, RSD, eluent type and concentration and detection technique of present work was compared with literature data in Table 8. In most cases, suggested method for Hg preconcentration with nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED is comparable to, or better than, some of the previously reported nano-Fe<sub>3</sub>O<sub>4</sub>-DOP-ED based modified adsorbents. Finally, the proposed method can be suggested as simple, sensitive, accurate and repeatable method for determination of Hg(II) after preconcentration. This new enrichment procedure can also be applied to some other metal ions.

# ACKNOWLEDGEMENT

The author wish to thank the Chemistry Department of Varamin Branch Islamic Azad University for financial support.

# REFERENCES

- Bulut V.N., Tufekci M., Duran C., Soylak M., Kantekin H., (2010). 3D Visualization of iron oxide nanoparticles in MRI of inflammatory. Clean-Soil Air Water, 38: 678-684.
- Graf D., Molitor F., Ensslin K., Stampfer C., Jungen A., Hierold C., (2007). Spatially Resolved Raman Spectroscopy Nano Lett., 7: 238-242.
- Ansari I.A., Dewani V.K., Khuhawar M.Y., (1999).

Flavonoid constituents of Chorizanthe diffusa with potential cancer chemopreventive activity. J. Chem. Soc. Pak., 21: 359-365.

- Margel S., Gura S., Bamnolker H., Nitzan B., Tennenbaum T., Bar-Toov B., Hinz M., Seliger H., in: Hafeli U., schutt W., Teller J., Zborowski M. (Eds.), (1997), Scientific and Clinical Applications of Magnetic Carriers, Plenum Press, New York, 481-494.
- Afkhami A., Moosavi R., (2010). Adsorptive removal of Congored, acarcinogenic textile dye, from aqueous solutions by maghemite nano particles. J. Hazard Mater., 174: 398-403.
- White B.R. Stackhouse B.T., Holocombe J.A., (2009).
  Magnetic γ-Fe<sub>2</sub>O<sub>3</sub> nanoparticles coated withpolyl-cysteine for chelation of As(III),Cu(II),Cd(II), Ni(II),Pb(II) and Zn(II). J. Hazard Mater., 161: 848-853.
- Tuutijarvi T., Lu J, Sillanpaa M., Chen G., (2009). As(V) adsorption on maghemitena- noparticles. J. Hazard Mater., 166: 1415-1420.
- Batterham G.J., Munksgaard N.C., Parry D.L., (1997). The effects of poverty on child health and development Sociology. J. Anal. At. Spectrom., 12: 1277-1282.
- Hummers W.S., Offeman R.E., (1958). Functionalized Graphene and Graphene Oxide: Materials Synthesis. J. Am. Chem. Soc., 80: 1339-1344.
- Camel V., (2003). Solid phase extraction of trace elements - CNSTN, Spectrochim. Acta Part B. 58, 1177-1182.
- Castillo M., Pina-Luis G., D'1az-Garcia M.E., Rivero I.A., Redalyc, (2005). Solid-Phase Organic Synthesis of Sensing Sorbent Materials. J. Braz. Chem. Soc., 16: 412-417.
- Eaton A.D., Clesceri L.S., Greenberg A.E., (1995). Standard Methods for the examination of water and waste water.19 th ed ,American Public Health Association, Washington, DC.
- Welcher F.J., Boschmann E., (1979). Organic Reagents for Copper, Krieger Huntington, New York.
- Marczenko Z., (1986). Separation and Spectrophotometric Determination of Elements, Ellis Horwood, London.
- Carasek E., Tonjes J.W., Scharf M., (2002). Solvent Microextraction: Theory and Practice. Quim

Nova. 25, 748-755.

- Ceccarini A., Cecchini I., Fuoco R., (2005). Flow injection micelle-mediated methodology for determination of lead. Microchem. J., 79: 21-29.
- Tuzen M., Soylak M., Citak D., Ferreira H.S., Korn M.G.A., Bezerra M.A., (2009). A preconcentration system for determination of copper and nickel in water and food samples employing flame atomic absorption spectrometry. J. Hazard. Mater., 162: 1041-1047.
- Choi Y.S., Choi H.S., (2003). High Functional Inorganic Polymers Containing Main Group 13 16 Elements in the Polymer Backbone Chain. Bull Korean Chem. Soc. 24: 222-229.
- Tuzen M., Narin I., Soylak M., Elci L., (2004). XAD-4/PAN Solid Phase Extraction System for Atomic Absorption Spectrometric Determinations of Some Trace Metals in Environmental Samples. Anal. Lett., 37: 473-480.
- Karousis N., Sandanayaka A.S.D., Hasobe T., Economopoulos S.P., Sarantopouloua E., Tagmatarchis N., (2011). The Role of Oxygen during Thermal Reduction of Graphene Oxide. J. Mater. Chem., 21: 109-118.
- Smith M.B., March J., (2001). March's advanced organic chemistry: reactions, mechanisms, and structure. New York: John Wiley & Sons Inc.
- Mermoux M., Chabre Y., Rousseau A., (1991). FTIR and 13CNMR study of graphite oxide. Carbon, 29(3): 469-74.
- Cataldo F., (2003). Structural analogies and differences between graphite oxide and C60 and C70 polymeric oxides (fullerene ozopolymers). Fuller Nanotub Car N., 11(1): 1-13.
- Tohidifar H., Moghimi A., Ayvazzadeh O., Eskandari, S., (2013). Determination of lead (II) in milk by flame atomic absorption spectrometry after solid phase extraction. Asia. J. Chem., 25(11): 5981-5987.
- Thistlethwaite P.J, Hook M.S., (2000). Diffuse reflectance Fourier transform infrared study of the adsorption of oleate/oleic acid onto titania. Langmuir, 16: 4993-5007.
- Taguchi H., Paal B., Armarego W.L.F., (1997). Glyceryl-ether monooxygenase [EC 1.14. 16.5], Part
  9. Stereospecificity of the oxygenase reaction. J. Chem. Soc. Perkin Trans, 1(3): 303-7.

- Tajodini N., Moghimi A., 2010. Simultaneous Preconcentration of Cadmium (II) and Lead (II) in Water Samples Using Resin Amberlite XAD-2 Functionalized with Nitroso R Salt and Determination with Atomic Absorption Spectrometry. Asia J. Chem., 22(5), 3349-3361.
- Stevens J., Crawford M., Robinson G., Roenneburg L., (2007). Automated post-collection concentration for purified preparative fractions via solid phase extraction. J. Chromatogr. A, 1142: 81-83.
- Su X.G., Wang M.J., Zhang Y.H., Zhang J.H., Zhang H.Q., Jin Q.H., (2003). Separation and preconcentration procedures for the determination of lead using spectrometric techniques: A review, Talanta, 59: 989-995.
- Soylak M., Karatepe A.U., Elci L., Dogan M., (2003). Column preconcentration/separation and atomic absorption spectrometric determinations of some heavy metals in table salt samples using amberlite XAD-1180. Tur. J. Chem., 27(2): 235-242.
- Szabo T., Berkesi O., Dekany I., (2005). DRIFT study of deuterium-exchanged graphite oxide. Carbon, 43: 3186-9.
- Moghimi A., (2014). Extraction of Ni(II) on micro crystalline naphthalene modified with organic-solution-processable functionalized nano graphene. Russ. J. Phys. Chem. A, 88 (7): 1177-1183.
- Moghimi A., Abdouss M., (2013). Extraction of Co(II) by Isocyanate Treated Graphite Oxides (iGOs) Adsorbed on Surfactant Coated C<sub>18</sub> before Determination by CVAAS. Int. J. Bio-Inorg. Hybd Nanomat., 2(1): 319-327.
- Moghimi A., Sabertehrani M., Waquif-Husain S., (2007). Preconcentration and determination of chromium species using octadecyl silica membrane disks and flame atomic absorption spectrometry. Chinese J. Chem., 25 (12): 1859-1865.
- Tarigh G.D., Shemirani F., (2013). Magneticmultiwall carbon nano tubenano composite as an adsorbent for preconcentration and determination of lead (II) and manganese (II) in variousmatrices. Talanta, 115: 744-750.
- Moghimi A., Poursharifi M.J., (2012). Perconcentration of Ni(II) from Sample Water by Modified Nano Fiber. Orient J. Chem., 28(1): 353-356.

Moghimi A., (2014). Separation and extraction of

Co(II) using magnetic chitosan nanoparticles grafted with  $\beta$ -cyclodextrin and determination by FAAS. Russ. J. Phys. Chem. A, 88(12): 2157-2164.

- Moghimi A., Siahkalrodi S.Y., (2013). Extraction and Determination of Pb(II) by Organic Functionalisation of Graphenes Adsorbed on Surfactant Coated  $C_{18}$  in Environmental Sample. J. Chem. Health Risk, 3(3): 1-12.
- Moghimi A., Yari M., (2014). Preconcentration of trace Ni(II) using C<sub>18</sub> disks nano graphene with amino propyltriethoxysilane (APTES). Merit Research Journal of Environmental Science and Toxicology (MRJEST). 2(5): 110-119.
- Moghimi A., Akbarieh S.P., (2014). Evaluation of Solid-phase Extraction Sorbent with Octadecanefunctionalized Nano Graphene (ODG) for the Preconcentration of Chromium Species in Water, Int. J. Sci. Res. in Knowledge, 2(1): 8-21.
- Moghimi A., (2013). Detection of trace amounts of Pb(II) by schiff base-chitosan-grafted multiwalled carbon nanotubes. Russ. J. Phys. Chem. A, 87 (7): 1203-1209.
- Moghimi A., Esfanjani S., Mazloomifar H., (2012). Extraction and Perconcentration of Lead (II) using Octadecyl Bonded Silica Cartridge and Its Determination by Flame Atomic Absorption Spectrometry. Asia. J. Chem., 24(10): 4697.
- Moghimi A., Abdouss M., Ghooshchi G., (2013). Preconcentration of Hg(II) by Graphene Oxide with Covalently Linked Porphyrin Adsorbed on Surfactant Coated C<sub>18</sub> before Determination by CVAAS. Int. J. Bio-Inorg. Hybd Nanomater., 2(2): 355-364.
- Moghimi A., Shabanzadeh M., (2012). Extraction and Determination of Trace Copper (II) Using Octadecyl Silica Membrane Disks Modified 1-(2-Pyridyl Azo) 2-Naphtol (Pan) in Water Samples and Paraffin-Embedded Tissues from Liver Loggerhead Turtles Specimens by FAAS. J. Chem.

Health Risk, 2(2): 7-12.

- Xie F., Lin X., Wu X., Xie Z., (2008). Solid phase extraction of lead (II), copper (II), cadmium (II) and nickel. Talanta, 74: 836-848.
- Goswami A., Singh A.K., Venkataramani B., (2003). Green Materials for Sustainable Water Remediation and Treatment. Talanta, 60: 1141-1153.
- Moghimi A., (2012). Adsorption and preconcentration of lead (II) by solid-phase extraction prior to determination by flame atomic absorption spectrophotometry. Aust. J. Basic & Appl. Sci., 6(3): 320-330.
- Moghimi A., (2007). Preconcentration and determination of trace amounts of heavy metals in water samples using membrane disk and flame atomic absorption spectrometry. Chin. J. Chem., 25(5): 640-644.
- McAllister M.J., Abdala A.A., Aksay I.A., Prudhomme R.K., (2007). Intercalation and stitching of graphite oxide with diaminoalkanes. Langmuir, 23: 10644-10649.
- Pourreza N., Mirzajani R., Kiasat A.R., Abdollahzadeh R., (2012). Simultaneous preconcentration of Cd(II), Cu(II) and Hg(II) on Nano Quim. Nova., 35: 1945-1956.
- Ekinci Dogan C., Akcin G., (2007). Solid Phase Extraction and Determination of Lead in Water Samples. Anal. Lett., 40: 2524-2529.
- Moghimi A., (2014). Solid phase extraction of Cu(II), Fe(III) and Pb(II) ions Nanographene with amino propyl triethoxysilane (APTES). IJBRITISH. 14(1): 1-9.
- Ghaedi M., Fathi M.R., Shokrollahi A., Shajarat F., (2006). The determination of Cu2+, Fe3+, Zn2+ and Hg<sup>2+</sup> in real samples. Anal. Lett., 39: 1171-1185.
- Ensafi A.A., Khayamian T., Karbasi M.H., (2003). Online preconcentration system for lead (II) determination in waste water. Anal. Sci., 19: 953-960.

lo

# **AUTHOR (S) BIOSKETCHES**

Ali Moghimi, Associate Professor, Department of Chemistry, Varamin (Pishva) Branch, Islamic Azad University, Varamin, Iran, *Email: alimoghimi@iauvaramin.ac.ir ; kamran9537@yahoo.com*