

Dispersive solid-phase extraction method with 2-aminopyridine/graphene oxide nano-plates of trace Co(II) in water samples

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ABSTRACT: A flame atomic absorption spectrophotometric (FAAS) method for the determination of trace Co(II) after adsorption of synthesis 2-aminopyridine/graphene oxide nano-plates has been developed. Applying a facile and accessible sample preparation method is highly important before the experimental analysis process. In the current study, a technique was introduced for preconcentration and analysis of trace amounts of Co(II) ions in water samples. The important parameters in the analytical procedure of Co(II) ions were optimized such as the pH of extraction, amount and type of elution solvent, time of extraction, and the effect of another ion. Analytical parameters such as the concentration factor, the limit of detection (LOD) of the technique, and relative standard deviation (RSD %) were achieved as 20, 8.0 $\mu\text{g L}^{-1}$, and 2.4%, respectively.

Keywords: *Co (II) ion, Extraction, Graphene oxide nano-plates, Heavy metal, Preconcentration of heavy metals, SPE.*

INTRODUCTION

The properties of nanomagnetic particles research on their functionality as magnetic carrier, also in medical imaging uses such as magnetic resonance imaging (MRI) It commonly, generated from industrial processes including electroplating, smelting, battery manufacturing, mining, metallurgy, and refining [2,3]. Cobalt might to pancreatic cancer, renal toxicity, and enhanced tumor growth. The US Environment Protection Agency (US-EPA) has classified it as one of the group B1 carcinogenic elements [4]. The maximum allowable total Cobalt in drinking water has been set at a concentra-

tion of 0.03 mg.L^{-1} by the World Health Organization (WHO) [5]. The direct determination of Cobalt in water samples by flame atomic absorption spectrometry (FAAS) is very difficult due to the low concentration of Cobalt ions and also interfering because of influences of the components of the matrix. Therefore, separation and preconcentration steps are often required to achieve accurate, sensitive and reliable results by FAAS. Several techniques including precipitation/co-precipitation [6], liquid-liquid extraction [7] cloud point extraction [8] and solid-phase extraction [9] (SPE) have been developed for the separation and preconcentration of trace Cobalt Among all methods, solid-phase extraction is the most common technique used for preconcentration

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of an analyte in environmental waters because of its advantages of high enrichment factor, high recovery, rapid phase separation, low cost, low consumption of organic solvents and the ability of combination with different detection techniques in the form of on-line or off-line mode [10, 11]. In recent years a new kind of SPE, magnetic solid-phase extraction (MSPE), has attracted so much attention. The MSPE is based on magnetic materials as sorbent, which can be isolated from the matrix quickly by using a magnet. Compared with other isolation methods, MSPE can improve the extraction efficiency and simplify the process of pre-processing. In recent year, the application of nanomaterial's such as activated carbon. Different methods, co-extractant ligands has attracted considerable attention [12-16]. However, the use of classical extraction methods for this purpose is usually for the Extraction and separation of Co ions have been suggested including liquid chromatography [17] supercritical fluid extraction [18,19], flotation [20], aggregate film formation [21-25], liquid membrane [26], column adsorption of pyrocatechol violet-Co complexes on activated carbon [27], ion pairing [28], ion pairing [29], preconcentration with yeast [30], and solid phase extraction using C18 cartridges and disks [31-33]. Consequently, in the current investigation, the researchers focused on the first application of 2-aminopyridine/graphene oxide nano-plates as a novel adsorbent for dispersive solid-phase and extraction of Co^{2+} wastewater samples before the flame atomic absorption spectrometry.

EXPERIMENTAL

Instrumentation

The determination of Co^{2+} by PG-990 flame atomic absorption spectrometer is equipped with HI-HCl which was done according to the recommendations of the manufacturers. Accordingly, the pH measurements were used by Sartorius model PB-11.

Materials

In the current investigation, the following materials have been applied for the experiments; graphene oxide nano-plates, hexahydrate Co (II) nitrate, thio-

semi-carbazide ligand, buffer, and nitric acid. Additionally, ethylenediamine, thio-semi-carbazide ligand ($\text{CH}_5\text{N}_3\text{S}$) was prepared from Darmstadt, Germany of Merck.

Synthesis GrO

To prepare GrO, a certain amount of graphite oxide powder (0.1 g) was placed in 100 mL of water and ethanol solution (50/50%, v/v) under ultrasonic power of 140 W for 2 h. The resulted powder was dried in a vacuum desiccator [31].

GrO functionalized with 2-aminopyridine

The 0.1 g of GrO powder was poured in 200 mL of deionized water, then 0.2 g of 2Ap was poured and the mixture was placed in a homogenizer for 30 min at 13000 rpm. Then, the 0.2 g KOH was added to the homogeneous mixture and was subjected to ultrasonic power of 140 W for 30 min. The precipitate was refluxed at 80 °C and then washed with water and ethanol and dried at 25 °C (Fig. 1) [32].

The initial experiment of Co^{2+} extraction for determining the suitable adsorbent

The procedure for extraction and recovery of Co ions by 2-aminopyridine/graphene oxide nano-plates is as follows; in the first step, 0.2 g of thiosemicarbazone ligand and 0.3 g of 2-aminopyridine/graphene oxide nano-plates and were dissolved in the little amount of acetone, and then it dried. Four 50-mL balloons were adopted, and then 0.05 g of 2-aminopyridine/graphene oxide nano-plates was poured into one of the balloons. Thereupon, a 1 ml buffer solution with a pH of 6.0 was added to balloons and a 2 ppm solution was prepared with an analyte. Four solutions were shaken at 25°C for 20 min, then these solutions for 15 min were centrifuged and the supernatant injected into an atomic absorption apparatus.

The effect of adsorbent amount for Co^{2+} extraction

Seven 2.0 ppm solutions with 50 mL of Co^{2+} were provided and poured into seven flasks. Seven solutions were adjusted at pH=6.0 (optimum pH) and different amounts of the adsorbent (0.005, 0.01, 0.03, 0.05, 0.07, 0.12, and 0.15) added to flasks. The mixtures were shaken for 20 min, and then the mixtures were

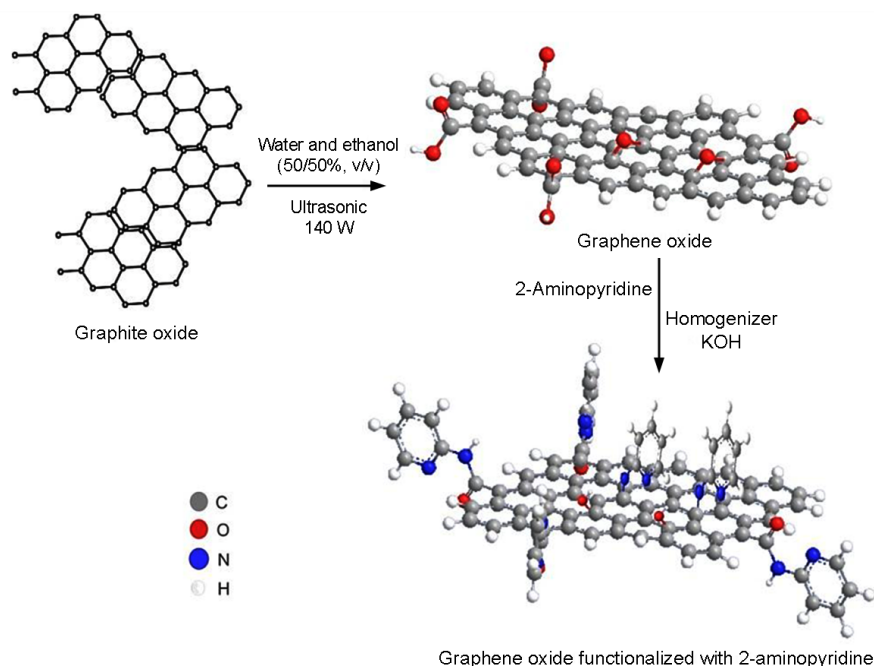


Fig. 1. Preparation of graphene oxide functionalized with 2-aminopyridine.

centrifuged and the top solution of the examine tube injected in flame atomic absorption spectrophotometry.

Application on real samples

Once the extraction method was performed by the adsorbent, optimal conditions were achieved for it, and Multiple real aqueous samples were investigated. The real samples were as follows; well and drinking water in Pishva Town were collected with temperatures of 20 and 22°C, pH=7.1, 7.3 in 23.8.95 at 9:45, 10:00, respectively. Finally, a fish farming sample was collected at, pH=6.20 in 23.8.95 at 11:20. First, these suitable bottles were provided for the sampling of samples. The bottles were washed first with ordinary and distilled water. The bottles dried completely, and the 'suitable' label was attached to each bottle. To collecting of water samples, the used containers sample dried and cleaned and they had already been washed. For the analysis of the samples in the first stage, colloidal and suspended particles were removed. To this aim, the water samples passed through 0.22 μm filters. Next, the volume of 100 mL of samples was poured into the sample container. The pH of samples was adjusted at 10 and then, graphene oxide nano-plates and ligand were added to samples. They stirred for 20 min

and the mixture was then centrifuged. Then, they were washed with HNO_3 0.1 M and were shaken again for 20 min. Finally, following the centrifugation of the mixture, absorption of Co ion was identified from the filtered solution by flame atomic absorption spectrophotometry. In the first step, the sample itself was injected into the apparatus without any Co ion, wherein water samples, the device displayed no absorption. To identify certain amounts of Co of the samples, the method of standard elevation was used. This stage was accomplished like the first step, the only discrepancy was that 0.5 mL of 200 ppm solution with Co^{2+} added to the water samples. Finally, the absorption of Co ion was identified from the filtered solution by flame atomic absorption spectrophotometry.

RESULT AND DISCUSSION

This section deals with the results of the research experiments. The results achieved in the experimental chapter, calibration curve, and the factors influencing the extraction (e.g. pH, temperature effect, time, etc.) of Co^{2+} ion by the 2-aminopyridine/graphene oxide nano-plates are discussed which are followed in the presentation of scientific justification and overall con-

clusion of the study.

Assessment of results of FT-IR spectrum of GrO and Gr2Ap

We synthesized Gr2Ap with the process of grafting of 2Ap on GrO. The recognition of the 2Ap group on the GrO surface via reaction of amination was demonstrated with the FT-IR test (Fig. 2c). Also seen, the spectrum of the GrO (Fig. 2a) shows two absorption bands in 1055 and 1401 cm^{-1} related to the C-O stretch bond from the carboxylic acid. Also, the band of tensile vibration bond of C-O-C was seen in the area of 1207 cm^{-1} . Moreover, a vibration band in 1629 cm^{-1} relating to the C=C bond belonging to unoxidized carbons. Finally, the strong band in 3443 cm^{-1} belongs to the O-H vibration [21]. In the FT-IR spectrum of 2Ap (Fig. 2b), the absorption band in 1630 cm^{-1} indexed to the C=N bond of the pyridine ring. The peak of 1558 cm^{-1} corresponds to the C=C of the aromatic ring. The peak of 1273 cm^{-1} ascribed to the C-N bond in the ring and also the peak of 3444 cm^{-1} related to N-H tensile vibrations [31].

Fig. 2c displays FT-IR spectra of Gr2Ap, this sample has an absorption band in 3448 cm^{-1} which originated from the group of OH from GrO. Two peaks at 2858 and 2923 cm^{-1} indicate asymmetric and symmetric vibrations of the CH_2 group [32-33]. The weak peak in 1162 cm^{-1} is related to the grafting of the amine group to GrO. Also, the peak intensity of the C-O at 1074 cm^{-1} decreased compared to the spectra of GrO. The peak in 1629 cm^{-1} is related to the C=N bond of the

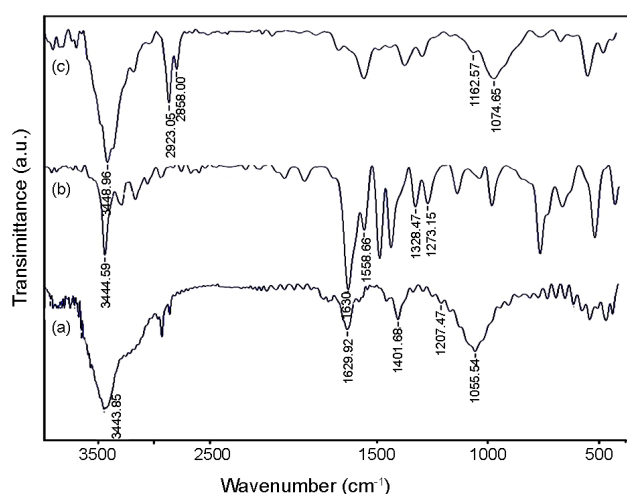


Fig. 2. FTIR spectra of GrO (a), 2Ap (b) and Gr2Ap (c).

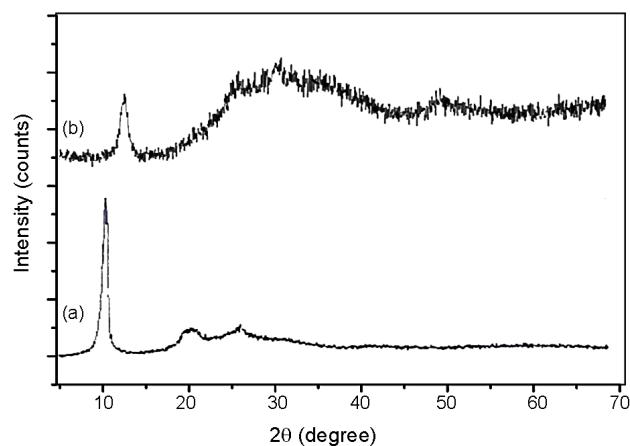


Fig. 3. XRD patterns of GrO (a) and Gr2Ap (b).

pyridine ring on the GrO surface. The band of 1207 cm^{-1} indexed to the bond of C-N in the ring of pyridine on GrO. The results obtained from the FT-IR analysis confirm the functionalization of GrO.

Fig. 3 represents X-ray diffraction patterns of synthesized GrO and Gr2Ap. Regarding the results, GrO shows a drastic peak at $2\theta=11.6^\circ$ demonstrating severe oxidation of graphite sheets and crystalline structure of GrO nano-plates (Fig. 3a) [34-36]. After functionalization with 2Ap, a new broad peak has appeared at 2θ ranging from 23 to 37, and also was decreased the intensity of the main peak (Fig. 3b). These results are respectively related to the disordering of the regularity of GrO nanoplates due to the incorporation of 2Ap groups between them as well as concurrent reduction of its oxygenated groups.

The morphology of GrO and Gr2Ap and the surface changes after functionalization was surveyed with SEM images. Fig. 4a is related to synthesized GrO through the improved Staudenmaier procedure. As is evident, the synthesized GrO has a finite number of layered structures with a flat surface. Fig. 4b1 implies that the GrO nano-plates retain their structure of layered while functionalizing process, but the structural regularity of the layers has been considerably decreased. Moreover, the 2Ap functional group is seen as spherical shapes on the GrO surface in sizes of 29-53 nm (Fig. 4b2). The width of GrO pores was distributed in the range of 2-10 nm according to the synthesis method in the reference, which can be classified as mesoporous.

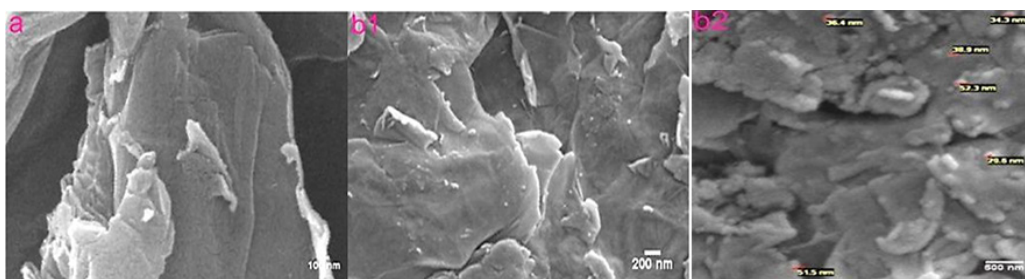


Fig. 4. SEM images of (a) GrO and (b1-2) Gr2Ap.

Investigation of the influential factors on Co^{2+} extraction

Study the effect of pH on Co^{2+} extraction

The results of this study are provided in Fig. 5. As the results in Fig. 5 indicate, at pH=6, Co adsorption was maximized, while at lowest and highest pHs, the extent of adsorption declines, inferring that at $pH < 6$ adsorptions of Co^{2+} ions cannot occur completely. As revealed in Fig. 5, to determine the amount of 2-aminopyridine/graphene oxide nano-plates required for effective removal of Co^{2+} , different amounts of the 2-aminopyridine/graphene oxide nano-plates (50 mg) for modification of 2-aminopyridine/graphene oxide nano-plates with fixes amount (3 mg) and its effect for the removal of Co^{2+} from 20 mL solutions of cobaltion (50 $\mu g/L$) were investigated.

Investigation of the effect of time on Co^{2+} extraction

Based on the results, the extent of absorption increases, and the Co ions present in the solution find more chance to be adsorbed in the adsorbent's sites. Therefore, the quantitative extraction of Co ion is possible for a period of longer than 20 min, and within durations longer than 20 min and more, the reaction hap-

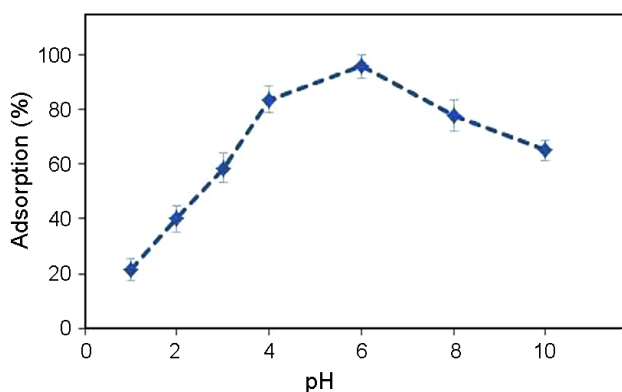


Fig. 5. The effect of pH in the Co^{2+} adsorption.

pens completely.

Survey of the effect of type of various desorption solvent for recovery of Co^{2+}

Based on the results (Table 1), NaOH cannot be used as appropriate desorption and these bases do not possess a complete detergency power. Therefore, mineral acids with determined concentrations, H_2SO_4 , and HNO_3 were applied. As shown in Table 1, the results of this table offer that all acids contain a good detergency power for Co^{2+} , but the recovery percentage of HNO_3 is higher than that of other acids. In an acidic environment, the possible deposits dissolved and recovery of these ions increased. However, the results obtained from nitric acid were better than H_2SO_4 , in that 0.1 M of solution washed 92.54% of the Co^{2+} ion adsorbent. So, for the rest of the experiments, nitric acid 0.10 M was used as the desorption solution.

Survey of optimization of the volume effect of desorption solvent for Co^{2+} recovery

After the investigation and choice of optimal desorption, the volume of solvent was investigated, with the

Table 1. Selection of the appropriate desorption for recovery of Co^{2+}

| Solvent | Recovery (%) |
|----------------|-------------------------|
| HNO_3 0.1M | 95.20(2.4) ^a |
| HNO_3 1M | 85.32(2.4) |
| HNO_3 3M | 75.55(2.4) |
| H_2SO_4 0.1M | 70.64(2.5) |
| H_2SO_4 1M | 70.70(2.5) |
| NaOH 0.1 M | 60.46(2.5) |

a) measurement RSD after three replications

Table 2. The optimum volume of the desorption solvent

| Solvent volume | Recovery (%) |
|----------------|-------------------------|
| 5 ml | 40.20(2.3) ^a |
| 7ml | 59.67(2.4) |
| 9 ml | 78.66(2.4) |
| 12 ml | 97.75(2.4) |
| 14 ml | 88.15(2.2) |
| 16ml | 89.96(2.5) |
| 20ml | 88.99(2.5) |

a) Measurement RSD following three replications

results which are shown in Table 2. The volume of 12 mL for HNO₃ was selected as the optimal volume for washing.

Investigation the effect of breakthrough volume

Following the optimization of the pH of the desorption solvent and sample solution, etc., to elute the Co²⁺ in the adsorbents, the maximum volume of the aqueous solution containing Co²⁺ should be measured. If the volume of the test solution to be less than the breakthrough volume, and passaging of that volume, all analytes are kept in the solid phase. The results in (Table 3) verify that up to 250 mL of ions are adsorbed by the nano adsorbents and if the sample volume is greater than this value, some of the Co²⁺ is not kept on the adsorbent and pass over the adsorbent with no inhibition. Also with definition by the concept of breakthrough volume, it can be reported that the breakthrough volume in the current study is 250 mL and if the sample solution volumes which includes Co²⁺ is over 250 mL,

Table 3. Investigations of the effect of solution volume in the sample

| V(ml) | Recovery (%) |
|-------|------------------------|
| 50 | 88.7(2.5) ^a |
| 100 | 78.2(2.4) |
| 150 | 77.7(2.0) |
| 250 | 75.5(2.9) |
| 350 | 62.4(2.0) |
| 500 | 35.9(2.5) |

a) Measurement RSD following three replications

Table 4. Measurement RSD following three replications

| Sample | Device response |
|--------|-------------------------|
| 1 | 0.023(2.8) ^a |
| 2 | 0.024(2.6) |
| 3 | 0.023(2.6) |
| 4 | 0.024(2.3) |

a) Measurement RSD following three replications

adsorption does not occur completely and hence if 250 mL of sample volume is passed over the adsorbent and then with 12 ml of the desorption solvent washed, the concentration factor could not be achieved as 30. This concept that the concentration of Co²⁺ in 7 mL of desorption solvent which was passed over the adsorbent grows by 20 times. Based on the related results (Table 3), the breakthrough volume calculations are as follows:

Concentration factor = breakthrough volume/the desorption solvent volume= 250/12=20

Determination of the blank standard deviation (S_b)

The accuracy or replicability of any method is the main factor to recognize its validity and reliability. To inquire about the method's replicability, the results data of the study of four blank solutions (deionized water) deposited in Table 4.

Based on the results achieved in Table 4, the blank standard deviation was obtained as follows; $S_b=0.0005$

Determination of the accuracy and RSD% of the method

This parameter was used to investigate the accuracy and proximity of the examined data. As shown in 2.4% standard deviation has been calculated for three tests and the relative standard deviation (RSD) achieved for three replications.

The linear range and a calibration curve of the method

To assess the linear range in the analysis method, a calibration curve should be plotted. This curve is not linear across all concentrations and different factors cause the calibration curve to the situation in the lin-

Table 5. The effect of interfering ions on the recovery of Co²⁺

| Ions | Added value (ppm) | Recovery percentage Co(II) |
|-------------------------------|-------------------|----------------------------|
| Na ⁺ | 200 | 2.5))92.92 |
| Zn ²⁺ | 5.0 | 2.4))91.34 |
| K ⁺ | 200 | 2.2))92.48 |
| Mg ²⁺ | 100 | 2.0))93.48 |
| Cu ²⁺ | 6.0 | 2.2))93.75 |
| Cl ⁻ | 400 | 2.6))92.17 |
| NO ₃ ⁻ | 317 | 2.0))92.27 |
| SO ₄ ²⁻ | 400 | 2.5))90.94 |

a) Measurement RSD following three replications

ear range and follow from Beyer Law. the calibration curve of the method is as conform and the line equation is $y=0.003x+0.075$ and $R^2=0.9994$.

Study of the effect of disturbances on the measurement of Co²⁺

A disturbing ion is an ion that causes a certain variation of over $\pm 5\%$ in the adsorption and recovery of Co²⁺. To study the effect of disturbance of other ions on Co²⁺ extraction, a certain quantity of interfering factors added to the initial solution, and the experiment was performed at breakthrough volume. Absorption of the recovered solution is analyzed with flame atomic absorption and then compared versus the solution absorption resulting from the sample recovery which lacks the interfering ion. As can be shown in Table 5, in the presence of external ions, Co recovery occurred with $\pm 5\%$ variations and the external ions had no particular effects on the analysis and cause no disturbance.

Determining the method's limit of detection

The lowest Co²⁺ concentration or weight in a sample that could be determined with a certain confidence level is called the limit of detection (LOD), which is defined as follows. The LOD of a method is a concentration of an analysis sample where the device response to concentration (which is significantly different from the response of the control sample) is defined as follows; the limit of detection is the lowest amount of Co²⁺, where the presented method can detect it. Based on the presented definition, LOD can be calculated by

the following relation;

$$\text{LOD} = \frac{3S_b}{m} \quad (1)$$

Where S_b and m are the standard deviations of the blank signal and the slope of the calibration curve, respectively. Based on the experimented, $S_b=0.0005$ and the slope of the calibration curve is 0.0002.

Therefore, LOD can be calculated at 8.0 ppb.

Investigation of the obtained results on real sample

The proposed method has been successfully applied to the determination of cobalt in three real drinking water samples (tap water of tehran, pure water and industrial wastewater sample of Charmshar Varamin. The results were shown in Table 6, together with results of a recovery test by added known amounts of silver in water sample. The results of this analysis are shown in Table 6. , the level of Co(II) ions was measured across different water and biological samples at 250 mL. As can be seen, in the water samples, in the Tap water sample of Tehran, on 20 Jan 2022, and industrial wastewater sample of Charmshar Varamin on 20 Jan 2022, there is a larger amount of Co²⁺ than in the experimented water samples. Although, in other samples, there is less Co(II) ions. Based on this, the performance and power of preconcentration and Co(II) ions measurement could be deduced.

A comparison between the current method and other methods

A comparison of this method with other methods veri-

Table 6. Determination of Co in the real sample

| Recovery | R.S.D.% (n=5) | This method ($\mu\text{g.g}^{-1}$)* | Reference method ($\mu\text{g.l}^{-1}$)* | Samples |
|----------|---------------|---------------------------------------|--|-----------------------|
| 98.3 | 2.4 | 10.8 | 11.2 | Tap water |
| 99.0 | 2.1 | 6.04 | 5.20 | Pure water |
| 95.7 | 2.4 | 19.2 | 18.9 | Industrial wastewater |

* Average of five determinations.

fied that the current method is more accurate, easiest, and faster as it had smaller relative standard deviation values in comparison with other methods [37-53]. The current method is one of the foremost systems for determining the very trace amounts of heavy metal ions including Co in aqueous samples. Another point in the usage of graphene oxide nano-plates adsorbent is that instead of using the proposed ligand, one can put other ligands on the adsorbent which to adsorb mineral ions, thereby measuring trace amounts of metal ions. A wide variety of ligands can be used given their properties, which act selective towards one or several ions and applying this set, preconcentration, and determination of cations can be carried out. Using flame atomic absorption and solid drop microextraction, single-drop liquid-liquid extraction, and homogeneous liquid-liquid extraction with other devices, one can determine trace amounts of Co^{2+} by this adsorbent and achieve a smaller limit of detection value.

CONCLUSIONS

In comparison with other procedures reported for measurement of Co (II), this method has considerable advantages that are easy and inexpensive and can be applied quickly for environmental aqueous samples. Furthermore, it minimizes the utilization of organic, toxic, and costly solvents. Moreover, the design and development of this procedure for separation, measurement, and preconcentration of Co^{2+} are essential considering its importance in various industries and the little concentration of Co^{2+} ion in most samples. Therefore, this research aims to present an effective, selective, cost-effective, and simple method for measurement of the level of Co (II) across different environmental aqueous samples (in this research, the limit of detection, the value of breakthrough volume,

and RSD has been obtained). This research indicated that the measurement of Co^{2+} occurs at an appropriate level without the interference of any other interfering factor and thus the current method can be applied easily in the measurement of the quantity of Co (II) in aqueous samples.

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REFERENCES

- [1] Deng, Y., Wang, C., Shen, X., Yang, W. (2005). Preparation, characterization, and application of multistimuli-responsive microspheres with fluorescence-labeled magnetic cores and thermoresponsive shells. *Chem. Eur. J.*, 116006–6013.
- [2] Modo, M., Hoehn, M., Bulte, Jeff W.M. (2005). Cellular MR imaging. *Mol. Imaging*, 4, 143–164.
- [3] Bulte, J.W.M. (2006). Intracellular endosomal magnetic labeling of cells. *Method Mol. Med.*, 124, 419–439.
- [4] Bruening, M.L., Mitchell, D.M., Bradshaw, J.S., Izatt, R.M., Bruening, R.L. (1991). Removal of cesium from alkaline waste solution: Part II – Column ion exchange study. *Anal. Chem.*, 63, 21-27.
- [5] Pourshamsi, T., Amri, F., Abniki, M. (2021). A comprehensive review on application of the syringe in liquid-and solid-phase microextraction methods. *J. Iran Chem. Soc.*, 18, 245–264.
- [6] Campderros, M.E., Acosta, A., Marchese, J.

- (1998). Selective separation of copper with Lix 864 in a hollow fiber module. *Talanta*, 47, 19-23.
- [7] Abniki, M., Moghimi, A., Azizinejad, F. (2020), Fabrication of bionanocomposite based on LDH using biopolymer of gum arabic and chitosan-coating for sustained drug-release. *J. Serb. Chem. Soc.*, 84,1-13.
- [8] Akama, Y., Ito, M., Tanaka, S. (2000). Selective separation of cadmium from cobalt, copper, iron (III) and zinc by water-based two-phase system of tetrabutylammonium bromide. *Talanta*, 52, 645-651.
- [9] Alexandrova, A., Arpadjan, S. (1993). Elimination of sulfide interference by sodium hypochlorite solution in the cold vapor atomic absorption spectrometric determination of mercury using tin(II) reduction in alkaline medium. *Analyst*, 118, 1309-1313.
- [10] Arpadjan, S., Vuchkova, L., Kostadinova, E. (1997). Study of the adsorption behavior of heavy metal ions on nanometer-size titanium dioxide with ICP-AES. *Analyst*, 122, 243-250.
- [11] Boudreau, S.P., Cooper, W.T. (1989). Analysis of thermally and chemically modified silica gels by heterogeneous gas-solid chromatography and infrared spectroscopy. *Anal. Chem.*, 61, 41-47.
- [12] Boll, I., Kramer, R., Brunner, J., Mokhir, A. (2005). Oligonucleotide-Templated Reactions for Sensing Nucleic Acids. *J. Am. Chem. Soc.*, 27, 7849-7855.
- [13] Brunner, J., Mokhir, A., Kramer, R. (2003). Copper(II) Quenched Oligonucleotide Probes for Fluorescent DNA Sensing. *J. Am. Chem. Soc.*, 125, 12410-12415.
- [14] Moghimi, A., Abniki, M., Khalaj, M., Qomi, M. (2021). Construction of modified nanotube carbon carboxyl by new method and application in dispersive solid phase extraction for preconcentration of Ni²⁺. *Rev. Roum. Chim.*, 66(6), 493-507.
- [15] Moghimi, A., Abniki, M. (2021). Dispersive Solid-Phase Extraction for Bromocresol Green Removal with β -Cyclodextrin Functionalized Magnetic Nanotubes. *Russ. J. Phys. Chem. B*, 15, S130-S139.
- [16] Moghimi, A., Abniki, M. (2022). Selective separation of copper with Lix 864 in a hollow fiber module), *J. Color Sci. Technol.*, 4, 301-315.
- [17] Caroli, C., Alimanti, A., Petrucci, F., Horvath, Z. (1991). Selective pre-concentration and solid phase extraction of mercury(II) from natural water by silica gel-loaded dithizone phases. *Anal. Chim. Acta*, 248, 241-247.
- [18] Choi, Y.S., Choi, H.S. (2003). Studies on Solvent Sublimation of Trace Heavy Metals by Continuous Flow System as Ternary Complexes of 1,10-Phenanthroline and Thiocyanate Ion. *Bull. Korean Chem. Soc.*, 24, 222-228.
- [19] Cuculic, V., Mlakar, M., Branica, M. (1997). Synergetic adsorption of copper(II) mixed ligand complexes onto the SEP-PAK C18 column. *Anal. Chim. Acta*, 339, 181-189.
- [20] Dadler, V., Lindoy, L.F., Sallin, D., Schlaepfer, C.W. (1987). Selective pre-concentration and solid phase extraction of mercury(II) from natural water by silica gel-loaded dithizone phases. *Aust. J. Chem.*, 40, 1557-1563.
- [21] Abniki, M., Moghimi, A. (2021). Synthesis of chitosan functionalized magnetic carbon nanotubes for dispersive solid-phase extraction of bromocresol green. *Micro & Nano Lett.*, 16, 455-462.
- [22] Moghimi, M., Qomi, M., Yari, M., Abniki, M. (2019). Solid phase extraction of Hg(II) in water samples by nano-Fe. *Int. J. Bio-Inorg. Hybr.*, 4(8), 168-159.
- [23] Tajodini, N., Moghimi, A., Karimnezhad, K. (2020). Separation of levodopa using Nanotubes Carbon modified Methyl Amine in biological samples and determination by UV-Vis Spectrophotometry. *J. Adv. Pharm. Educ. Res.*, 10(S4), 153-163.
- [24] Grote, M., Kettrup, A. (1985). Liquid-liquid extraction of noble metals by formazans: Analytical Applications of Silver Extraction by ortho-Substituted Formazans. *Anal. Chim. Acta*, 175, 239-244.
- [25] Hummers, W.S., Offeman, RE. (1958). Preparation of graphitic oxide. *J. Am. Chem. Soc.*, 80, 1339-1344.
- [26] Krueger, C.J., Fild, J.A. (1995). Method for the analysis of triadimefon and ethofumesate from dislodgeable foliar residues on turfgrass by

- solid-phase extraction and in-vial elution. *Anal. Chem.*, 67, 3363-3369.
- [27] Kvittek, R.J., Evans, J.F., Carr, P.W. (1982). Denaturation of purple membranes at the air/water interface studied by SEM. *Anal. Chim. Acta*, 144, 93-98.
- [28] Leyden, D.E., Luttrell, G.H., Nonidez, W.K., Werho, D.B. (1976). Adsorption of Co(II) and Cu(II) on silica gel surface modified with pyridinium ion from acetone and ethanol solutions. *Anal. Chem.*, 48, 67-72.
- [29] Leyden, D.E., Luttrell, G.H., Sloan, A.E., DeAngelis, N.J. (1976). Automated separation and pre-concentration of copper(II) from natural waters using a column treatment. *Anal. Chim. Acta*, 84, 97-102.
- [30] Moghimi, A., Abniki, M. (2021). The Dispersive Solid-Phase Extraction of Fluoxetine Drug from Biological Samples by the Amine-Functionalized Carbon Nanotubes with HPLC Method. *Chem. Methodol.*, 5(3), 250-258.
- [31] Liu, J., Wang, Y., Xu, S., Sun, D.D. (2010). Synthesis of graphene soluble in organic solvents by simultaneous ether-functionalization with octadecane groups and reduction. *Mater. Lett.*, 64, 2236-2239.
- [32] Salehi, N., Moghimi, A., Shahbazi, H. (2021). Magnetic nanobiosorbent (MG-Chi/Fe₃O₄) for dispersive solid-phase extraction of Cu(II), Pb(II), and Cd(II) followed by flame atomic absorption spectrometry determination. *IET Nanobiotechnol.*, 1-10.
- [33] Mahmoud, M.E. (1997). Silica-immobilized formylsalicylic acid as a selective phase for the extraction of iron(III). *Talanta*, 44, 15-21.
- [34] Mahmoud, M.E., Soliman, E.M. (1997). Study of the selective extraction of iron(III) by silica-immobilized 5-formyl-3-aryloxy-salicylic acid derivatives. *Talanta*, 44, 1063-1071.
- [35] Moghimi, A., Abniki, M. (2021). Preconcentration and Separation of Ultra-Trace Cu(II) with Disks of Octadecyl Silica Membrane Modified Nano-Fe₃O₄-Encapsulated-Dioctyl Phthalate and Linked-Diethylenetriamine. *Adv. J. Chem. A*, 4(2), 78-86.
- [36] Moghimi, A., Alborji, A., Qomi, M., Anaraki Ardakani, H. (2020). An alternative method of extracting of trace Hg(II) in water samples using Fe₃O₄@quillaja Sapogenin on Ambersorb 572 and determination by CVAAS. *Arch. Pharm. Pract.*, 11(S1).
- [37] McAllister, M.J., Abdala, A.A., McAllister, M.J., Aksay, I.A., Prudhomme, R.K. (2007). *Langmuir*, 23, 10644-9.
- [38] Moghimi, A., Tajodini, N., Karimnezhad, K. (2021). Alternative Method of Extraction of ultra-trace Co(II) with disks of octadecyl silica membrane modified nano-Fe₃O₄-encapsulated-dioctyl phthalate and linked-diethylenetriamine. *Eur. J. Mol. Clin. Med.*, 8(1), 2080-2088.
- [39] Moghimi, A., Ghiasi, R., Abedin, A.R., Ghamamy, S. (2009). Solid phase extraction of Cd(II) using mesoporous organosilicas and determination by FAAS. *African J. Pure Appl. Chem.*, 3(3), 051-059.
- [40] Moghimi, A., Tajodini, N. (2010). Preconcentration of Copper(II) in Water Samples using Polyurethane Foam/2-(6'-Ethyl-2'-benzothiazolylazo) chromotropic Acid. *Asian J. Chem.*, 22(5), 3325-3334.
- [41] Abniki, M., Moghimi, A. (2022), Removal of Cd (II) ions from water solutions using dispersive solid-phase extraction method with 2-amino-pyridine/graphene oxide nano-plates. *Curr. Anal. Chem.*, Impress.
- [42] Moghimi, A. (2007). Preconcentration and Determination of Trace Amounts of Heavy Metals in Water Samples using Membrane Disk and Flame Atomic Absorption Spectrometry. *Chin. J. Chem.*, 25(10), 640-645.
- [43] Moghimi, A. (2008). Preconcentration of Copper(II) using Mesoporous Organo-Silicas and Determination by Flame Atomic Absorption Spectrometry. *J. Korean Chem. Soc.*, 52(2), 155-163.
- [44] Abniki, M., Moghimi, A., Azizinejad, F. (2021). Synthesis of calcium-layered double hydroxide based nanohybrid for controlled release of an anti-inflammatory drug. *J. Chin. Chem. Soc.*, 68(2), 343-352.
- [45] Moghimi, A., Tehrani, M.S., Waqif Husain, S. (2006). Preconcentration and Determination of

- Copper(II) using Octadecyl Silica Membrane Disks Modified by 1,5-Diphenylcarbazide and Flame Atomic Absorption Spectrometry. *Mate. Sci. Res. India*, 3(1a), 27-32.
- [46] Moghimi, A., Abdouss, M. (2012). Preconcentration of Ni(II) from sample water by modified poly (ethylene terephthalate)-grafted-acrylic acid/acryl amide fiber. *Afr. J. Pure Appl. Chem.*, 6(8), 110-118.
- [47] Moghimi, A. (2014). Separation and extraction of Co(II) using magnetic chitosan nanoparticles grafted with β -cyclodextrin and determination by FAAS. *Russ. J. Phys. Chem. A*, 88(12), 2157-2164.
- [48] Moghimi, A., Yari, M. (2019). Review of procedures involving separation and Solid Phase Extraction for the determination of cadmium using spectrometric techniques. *J. Chem. Rev.*, 1(1), 1-18
- [49] Nambiar, D.C., Patil, N.N., Shinde, V.M. (1998). Liquid-liquid extraction of mercury(II) with triphenylphosphine sulphide: Application to medicinal and environmental samples. *Fresenius J. Anal. Chem.*, 360, 205-212.
- [50] Narin, I., Soylak, M., Elic, L., Dogan, M. (2000). An Evaluation of Loading Rate of Dust, Pb, Cd, and Ni and Metals Mass Concentration in the Settled Surface Dust in Domestic Houses and Factors Affecting Them. *Talanta*, 52, 1041-1047.
- [51] Shojai, M., Moghimi, A., Reza Asghari, R. (2015). Preconcentration of Pb(II) on Micro Crystalline Naphthalene Modified with Organic Solution Processable Functionalized-Nano Graphene. *Elixir Appl. Chem.*, 82, 32605-32609.
- [52] Tajodini, N., Moghimi, A. (2010). Preconcentration and Determination of Ultra Trace Cobalt(II) in Water Samples using Co(II)-Imprinted Diazoaminobenzene- Vinylpyridine Copolymers. *Asian J.Chem.*, 22(5), 3335-3344.
- [53] Abniki, M., Moghimi, A. (2022). Removal of Cd (II) ions from water solutions using dispersive solid-phase extraction method with 2-aminopyridine/graphene oxide nano-plates, *Current Analytical Chemistry*, 18(10).

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